



**QUEEN'S
UNIVERSITY
BELFAST**

Systems Analysis of the Liver Transcriptome in Adult Male Zebrafish Exposed to the Plasticizer (2-Ethylhexyl) Phthalate (DEHP)

Huff, M., da Silveira, W., Carnevali, O., Renaud, L., & Hardiman, G. (2018). Systems Analysis of the Liver Transcriptome in Adult Male Zebrafish Exposed to the Plasticizer (2-Ethylhexyl) Phthalate (DEHP). *Nature Scientific Reports*, 8(1), 1-17. Article 2118. <https://doi.org/10.1038/s41598-018-20266-8>, <https://doi.org/10.1038/s41598-018-20266-8>

Published in:
Nature Scientific Reports

Document Version:
Publisher's PDF, also known as Version of record

Queen's University Belfast - Research Portal:
[Link to publication record in Queen's University Belfast Research Portal](#)

Publisher rights

Copyright 2018 the authors.

This is an open access article published under a Creative Commons Attribution License (<https://creativecommons.org/licenses/by/4.0/>), which permits unrestricted use, distribution and reproduction in any medium, provided the author and source are cited.

General rights

Copyright for the publications made accessible via the Queen's University Belfast Research Portal is retained by the author(s) and / or other copyright owners and it is a condition of accessing these publications that users recognise and abide by the legal requirements associated with these rights.

Take down policy

The Research Portal is Queen's institutional repository that provides access to Queen's research output. Every effort has been made to ensure that content in the Research Portal does not infringe any person's rights, or applicable UK laws. If you discover content in the Research Portal that you believe breaches copyright or violates any law, please contact openaccess@qub.ac.uk.

Open Access

This research has been made openly available by Queen's academics and its Open Research team. We would love to hear how access to this research benefits you. – Share your feedback with us: <http://go.qub.ac.uk/oa-feedback>

SCIENTIFIC REPORTS



OPEN

Systems Analysis of the Liver Transcriptome in Adult Male Zebrafish Exposed to the Plasticizer (2-Ethylhexyl) Phthalate (DEHP)

Matthew Huff^{1,2}, Willian A. da Silveira^{1,3}, Oliana Carnevali⁴, Ludivine Renaud⁵ & Gary Hardiman^{1,5,6,7,8}

The organic compound diethylhexyl phthalate (DEHP) represents a high production volume chemical found in cosmetics, personal care products, laundry detergents, and household items. DEHP, along with other phthalates causes endocrine disruption in males. Exposure to endocrine disrupting chemicals has been linked to the development of several adverse health outcomes with apical end points including Non-Alcoholic Fatty Liver Disease (NAFLD). This study examined the adult male zebrafish (*Danio rerio*) transcriptome after exposure to environmental levels of DEHP and 17 α -ethinylestradiol (EE2) using both DNA microarray and RNA-sequencing technologies. Our results show that exposure to DEHP is associated with differentially expressed (DE) transcripts associated with the disruption of metabolic processes in the liver, including perturbation of five biological pathways: 'FOXA2 and FOXA3 transcription factor networks', 'Metabolic pathways', 'metabolism of amino acids and derivatives', 'metabolism of lipids and lipoproteins', and 'fatty acid, triacylglycerol, and ketone body metabolism'. DE transcripts unique to DEHP exposure, not observed with EE2 (i.e. non-estrogenic effects) exhibited a signature related to the regulation of transcription and translation, and ruffle assembly and organization. Collectively our results indicate that exposure to low DEHP levels modulates the expression of liver genes related to fatty acid metabolism and the development of NAFLD.

Endocrine Disrupting Compounds (EDCs) are ubiquitous chemical compounds used in numerous consumer products as plasticizers or flame retardants that have been shown to have unforeseen impacts on the ecosystem¹ and human health². As their chemical structures are similar to the structures of natural hormones³, EDCs are able to bind and activate many receptors, including nuclear hormone receptors^{4,5}, and disrupt the endocrine system⁶. In light of these recent studies, EDCs are now considered chemicals of emerging concern⁷, and it is critical to fully understand the impact they might have on the health of the ecosystem and mankind at environmental levels.

A specific subset of EDCs, the xenoestrogens (XEs), are able to mimic 17 β -estradiol, a natural female estrogen⁸. The effects of these XEs on specific cell types, including prostate, fibroblast, and neural cells, have recently been characterized^{9,10}. However, their impact on the liver, the main site of detoxification and metabolism of xenobiotics¹¹, has not been as well defined. Commonly used as a plasticizer of polyvinyl chloride (PVC), di(2-ethylhexyl) phthalate (DEHP), a XE gives PVC flexibility, strength and bendability, and is currently the only phthalate used in PVC medical devices^{12,13}, the most highly DEHP exposed patients are neonates in the neonatal intensive care unit¹⁴. It is also present in office supplies and dust¹⁵ (notebooks, report covers and sheet protectors) and children's toys¹⁶.

¹MUSC Bioinformatics, Center for Genomics Medicine, Medical University of South Carolina, Charleston, SC, USA.

²MS in Biomedical Sciences Program, Medical University of South Carolina, Charleston, SC, USA. ³Department of Pathology and Laboratory Medicine, Medical University of South Carolina, Charleston, SC, USA. ⁴Dipartimento di Scienze della Vita e dell'Ambiente, Università Politecnica delle Marche, 60131, Ancona, Italy. ⁵Department of Medicine, Nephrology, Medical University of South Carolina, Charleston, SC, USA. ⁶Department of Medicine, University of California San Diego, La Jolla, CA, USA. ⁷Department of Public Health Sciences, Medical University of South Carolina, Charleston, SC, USA. ⁸Laboratory for Marine Systems Biology, Hollings Marine Laboratory, Charleston, SC, USA. Correspondence and requests for materials should be addressed to G.H. (email: hardiman@musc.edu)

⁹Department of Pathology and Laboratory Medicine, Medical University of South Carolina, Charleston, SC, USA. ¹⁰Department of Pathology and Laboratory Medicine, Medical University of South Carolina, Charleston, SC, USA. ¹¹Department of Medicine, Nephrology, Medical University of South Carolina, Charleston, SC, USA. ¹²Department of Medicine, University of California San Diego, La Jolla, CA, USA. ¹³Department of Public Health Sciences, Medical University of South Carolina, Charleston, SC, USA. ¹⁴Laboratory for Marine Systems Biology, Hollings Marine Laboratory, Charleston, SC, USA. Correspondence and requests for materials should be addressed to G.H. (email: hardiman@musc.edu)

Two recent studies suggested that DEHP may cause lipid accumulation and nonalcoholic fatty liver disease (NAFLD) by promoting PPAR α and sterol regulator element-binding protein 1c (SREBP-1c) expression^{17,18}. Furthermore, exposure to DEHP has been found to disrupt the insulin signaling pathway in rats and the human L-02 cell line through activation of PPAR γ ⁴, reducing the ability of the liver to maintain glucose homeostasis, leading to insulin resistance.

In this study, we examined the effect of exposure to 5.8 nM of DEHP on the liver transcriptome, a concentration that is relevant to observed environmental levels¹⁹. In the United States, concentrations of DEHP in wastewater derived from Oakland, CA ranged from 2.53 to 6913.2 nM²⁰. Studies in model organisms indicated that DEHP exposures ranging from 0.01 to 25.6 nM are sufficient to negatively affect animal growth and reproduction^{21,22}.

To examine the effects of DEHP exposure on the adult male hepatic transcriptome, we exploited the zebrafish model (*Danio rerio*) and undertook a systems level analysis. We performed microarray and RNA sequencing analyses, and considered the data in the context of the recently described Adverse Outcomes Framework (AOF)²³. This approach defines an adverse outcome as the end result of a causal series of events punctuated by “key events” (KE), the results of a molecular initiating event (MIE), a molecular interaction between a chemical stressor and a target biomolecule that alters gene expression. In parallel we assessed the effect of the E2 analogue 17 α -ethinylestradiol (EE2) with the objective of deciphering the estrogenic and non-estrogenic effects of DEHP. This study is the first to use deep transcriptome profiling to explore the effects of exposure to DEHP on the zebrafish liver.

Results

Molecular changes in DEHP exposed livers revealed altered translational regulation. To examine the effect of 21-day exposure to DEHP (5.8 nM) or EE2 (0.65 nM) on the liver of adult male zebrafish, we carried out a microarray experiment and analyzed the transcriptome of exposed and control fish. Unique array probes were ranked using an interest statistic described previously, which reflects the understanding that the gene with a greater absolute fold change is potentially more interesting^{24–26}. Of the top 3,000 ranked DE transcripts, 1,454 mRNAs were shared amongst the EE2 and DEHP exposures. EE2 and DEHP exposures altered the expression of 1,090 and 1,072 unique mRNAs respectively (Fig. 1A). Next, we analyzed the top ranked 3,000 DE transcripts using the Gene Ontology (GO) enrichment analysis and visualization tool (GORilla)²⁷. DEHP exposure enriched a number of carbohydrate metabolism processes, including chitin metabolic and chitin catabolic processes ($q = 4.46E-05$ and $8.91E-05$ respectively), amino sugar catabolic process ($q = 1.16E-04$) and ion transport ($q = 3.63E-02$) (Table 1, DEHP Total and Fig. 1B). Exposure to EE2 (Table 1, EE2 Total) was associated with enrichment in biological processes such as organic acid metabolic process ($q = 9.31E-03$), carboxylic acid metabolic processes ($q = 1.14E-02$), and oxoacid metabolic process ($q = 1.16E-02$), among other terms associated with lipid and fatty acid metabolism (Fig. 1C). We next performed enrichment analysis using DE gene signatures that were unique for DEHP and observed significant enrichment in terms related to regulation of translational initiation ($q = 1.65E-01$), negative regulation of translation ($q = 2.48E-01$), ruffle organization ($q = 1.93E-01$) and ruffle assembly ($q = 2.14E-01$) (Table 1, DEHP Unique and Fig. 1D). We also performed enrichment analysis using DE gene signatures unique to EE2, but found no significant enrichments (data not shown).

Functional enrichment analysis of human orthologs revealed altered transcriptional and translational pathways. We mapped zebrafish genes of interest to their human orthologs using Ensembl^{28,29} genes to take advantage of the more extensive functional and non-inferred electronic annotations available based on the human genome, as illustrated in Fig. 2⁶. Using the assigned human orthologs, we performed enrichment analysis with the “Transcriptome, ontology, phenotype, proteome, and pharmacome annotations” based gene list functional enrichment analysis (ToppFun) tool and observed that exposure to DEHP and EE2 both had a significant impact on metabolism (Table 2, Supplementary Tables S1–S4), including the organonitrogen compound biosynthetic process (DEHP Total, $q = 3.87E-16$) and oxoacid metabolic process (EE2 Total, $q = 1.03E-14$). Furthermore, exposures to EE2 and DEHP were both linked with significant enrichment relating to apoptotic processes (DEHP Total, programmed cell death, $q = 2.76E-10$ and EE2 Total, apoptotic process, $q = 1.38E-11$), while DE genes unique to DEHP were associated with enrichment in cell cycle-related processes (DEHP Total, $q = 5.23E-11$). DEHP exposure resulted in a unique enrichment signature related to RNA processing (DEHP Unique, $q = 1.80E-04$) and ribonucleoprotein complex biogenesis (DEHP Unique, $q = 1.19E-03$). In terms of co-expressed genes, we identified an overlap in the DE genes of interest to this study with genes that are up-regulated in the zebrafish crash & burn (crb) mutant for the bmyb gene (EE2 Total, $q = 6.07E-21$) and genes upregulated in hepatoblastoma (liver cancer cell) samples (DEHP Total, $q = 2.26E-24$). In DE genes expressed only after exposure to DEHP, we found commonality with genes up-regulated by activation of the mammalian target of rapamycin complex 1 (mTORC1) pathway (Table 2, DEHP Unique $q = 8.83E-06$, Supplementary Figure S1 and Supplementary Table S5).

Using ToppFun’s pathway analysis module to examine exposure to DEHP, we identified five pathways of interest (Table 3 and Supplementary Table S6): FOXA2 and FOXA3 transcription factor networks (Pathway ID = 137911), Metabolic pathways (132956), metabolism of amino acids and derivatives (1270158), metabolism of lipids and lipoproteins (1270001), and fatty acid, triacylglycerol, and ketone body metabolism (1270010). Our data suggests that DEHP influences expression of genes associated with metabolism, in particular the metabolism of lipids and fatty acids.

RNA-Seq analysis identified changes in cellular response and translation. High-throughput RNA-Seq was carried out to further explore transcriptomic changes in response to DEHP exposure using an advanced analytical pipeline we have recently described³⁰. Based on DESeq2 analysis, genes with an adjusted p -value of less than or equal to 0.4 were selected^{31–33}. We found that EE2 and DEHP altered the expression of 326

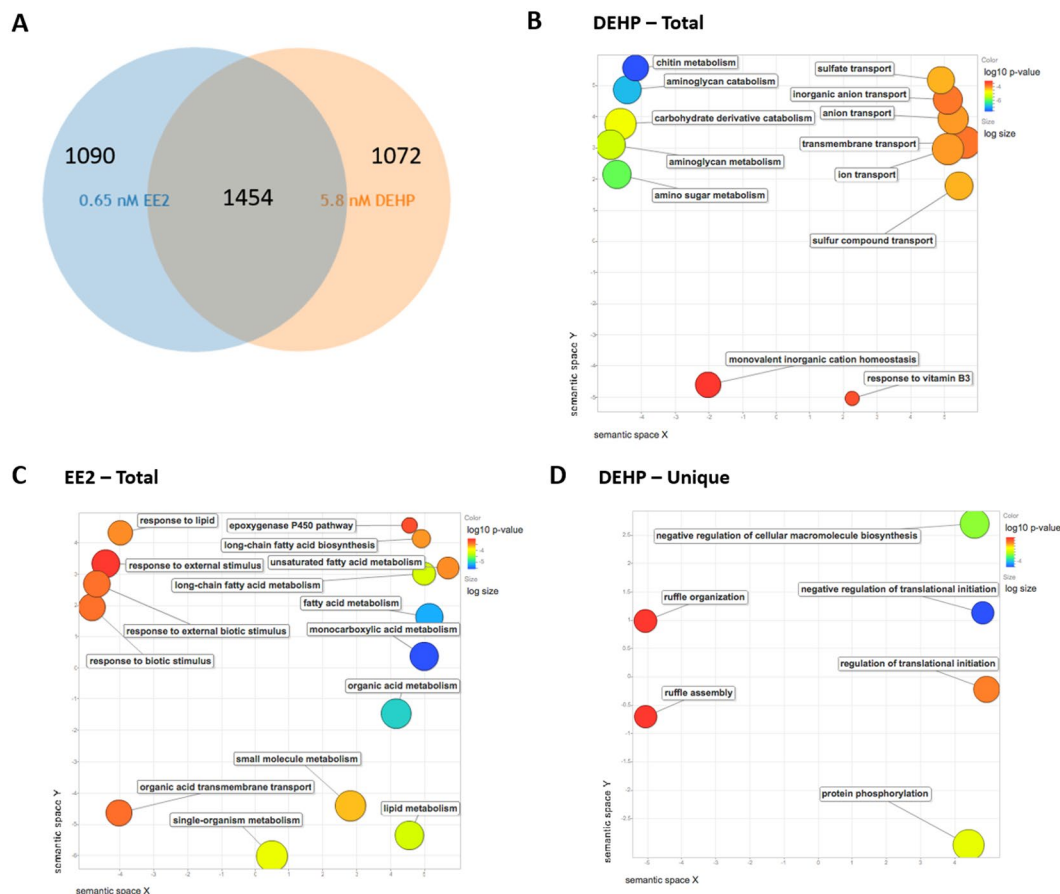


Figure 1. Functional Analyses. **(A)** Overlap of the top 3,000 ranked DE expressed liver transcripts from 5.8 nM DEHP and 0.65 nM EE2 exposed adult male zebrafish relative to control fish as determined by microarray analysis. **(B–D)** Gene Ontology Biological Process analyses: Scatterplots shows the cluster representatives (i.e. terms remaining after the redundancy reduction) in a two dimensional space derived by applying multidimensional scaling to a matrix of the GO terms' semantic similarities. Bubble color indicates the p-value (legend in upper right-hand corner); size indicates the frequency of the GO term in the underlying GOA database (bubbles of more general terms are larger). GO BP analysis of DE genes in **(B)** DEHP and **(C)** EE2 exposed livers. **(D)** GO BP analysis of DE genes unique to DEHP (not DE in EE2 exposed).

total genes; 66 genes were differentially expressed upon exposure to both DEHP and EE2, whereas exposures to EE2 and DEHP altered expression in 40 and 220 unique transcripts respectively (Fig. 3A).

Using these significant DE transcripts, we performed GOrilla analysis and found that exposures to both DEHP and EE2 were associated with enrichment of metabolic processes, particularly those relating to lipid metabolism (Table 4 and Fig. 3B,C), including lipid biosynthetic (DEHP Total, $q = 7.01E-08$ and EE2 Total, $q = 8.58E-07$) and metabolic processes (DEHP Total, $q = 1.05E-08$, EE2 Total, $q = 2.58E-04$) (all q -values are Bonferroni adjusted). Exposure to DEHP (DEHP Total, Table 4 and Fig. 3B) was associated with enrichments relating to translation ($q = 6.09E-14$), including cytoplasmic translation ($q = 5.19E-04$), and cellular responses to hypoxia ($q = 1.24E-02$) and response to estrogen ($q = 2.94E-04$). For exposure to EE2 (EE2 Total, Table 4 and Fig. 3C), we found unique enrichments related to the insulin signaling pathway ($q = 1.03E-01$) and the triglyceride metabolic pathway ($q = 2.05E-02$); the latter is supported with enrichments in triglyceride catabolism ($q = 9.03E-02$) and triglyceride-rich lipoprotein particle remodeling ($q = 6.13E-02$). Analysis of only the genes associated with exposure to DEHP (DEHP Unique, Table 4 and Fig. 3D) returned significant enrichment relating to translation ($q = 3.10E-16$), fatty acid biosynthesis ($q = 1.26E-15$), and the protein metabolic processes ($q = 1.29E-02$). No significant enrichments were detected with analysis of DE genes unique to EE2.

Functional enrichment analysis of RNA-seq data projected onto human orthologs identified dysregulation of translational and insulin-related pathways. We performed a functional enrichment analysis of the human orthologs using ToppFun and found that exposure to DEHP and EE2 is related to enrichment in metabolic processes (Table 5). In particular, EE2 enriched several lipid-related processes, including lipid biosynthesis ($q = 3.87E-09$) and fatty acid metabolism ($q = 7.42E-08$) (Table 5, EE2 Total; Supplementary Table S7). DEHP affected biological processes (Table 5, DEHP Total; Supplementary Table S8), such as co-translational protein localization to membranes ($q = 1.01E-19$), nuclear-transcribed mRNA catabolic process ($q = 1.79E-16$) and translational initiation ($q = 2.84E-15$). Analysis of genes unique to DEHP exposure

GO Term	q-value
DEHP-Total	
chitin metabolic process	4.46E-05
chitin catabolic process	8.91E-05
amino sugar catabolic process	1.16E-04
glucosamine-containing compound metabolic process	1.39E-04
glucosamine-containing compound catabolic process	1.74E-04
aminoglycan catabolic process	2.31E-04
amino sugar metabolic process	8.38E-04
aminoglycan metabolic process	3.38E-03
carbohydrate derivative catabolic process	8.12E-03
sulfur compound transport	2.82E-02
sulfate transport	3.10E-02
anion transport	3.44E-02
ion transport	3.63E-02
inorganic anion transport	5.55E-02
transmembrane transport	5.71E-02
response to vitamin B3	1.28E-01
organonitrogen compound catabolic process	1.69E-01
monovalent inorganic cation homeostasis	2.07E-01
EE2-Total	
organic acid metabolic process	9.31E-03
monocarboxylic acid metabolic process	1.12E-02
carboxylic acid metabolic process	1.14E-02
oxoacid metabolic process	1.16E-02
fatty acid metabolic process	1.53E-02
long-chain fatty acid metabolic process	4.58E-02
lipid metabolic process	5.10E-02
single-organism metabolic process	5.51E-02
small molecule metabolic process	1.12E-01
unsaturated fatty acid metabolic process	1.25E-01
carboxylic acid transmembrane transport	1.30E-01
response to lipid	1.35E-01
long-chain fatty acid biosynthetic process	1.38E-01
organic acid transmembrane transport	1.39E-01
response to biotic stimulus	1.44E-01
response to external biotic stimulus	1.56E-01
epoxygenase P450 pathway	1.87E-01
monounsaturated fatty acid metabolic process	2.14E-01
response to external stimulus	2.23E-01
monounsaturated fatty acid biosynthetic process	2.26E-01
DEHP-Unique	
negative regulation of cellular biosynthetic process	1.35E-01
negative regulation of gene expression	1.44E-01
negative regulation of biosynthetic process	1.62E-01
regulation of translational initiation	1.65E-01
negative regulation of cellular macromolecule biosynthetic process	1.74E-01
negative regulation of translational intiation	1.87E-01
protein phosphorylation	1.91E-01
ruffle organization	1.93E-01
ruffle assembly	2.14E-01
negative regulation of translation	2.48E-01
negative regulation of macromolecule biosynthetic process	2.62E-01
negative regulation of cellular amide metabolic process	2.71E-01

Table 1. GOrilla functional enrichment analysis. The top 3,000 ranked DE expressed liver transcripts as determined by microarray analysis from 5.8 nM DEHP and 0.65 nM EE2 exposed adult male zebrafish relative to control fish were used to search for enriched GO: Biological Process terms. The most significant terms for the DEHP and EE2 exposures, and those unique to DEHP (i.e. not present in in EE2 exposed) are presented.

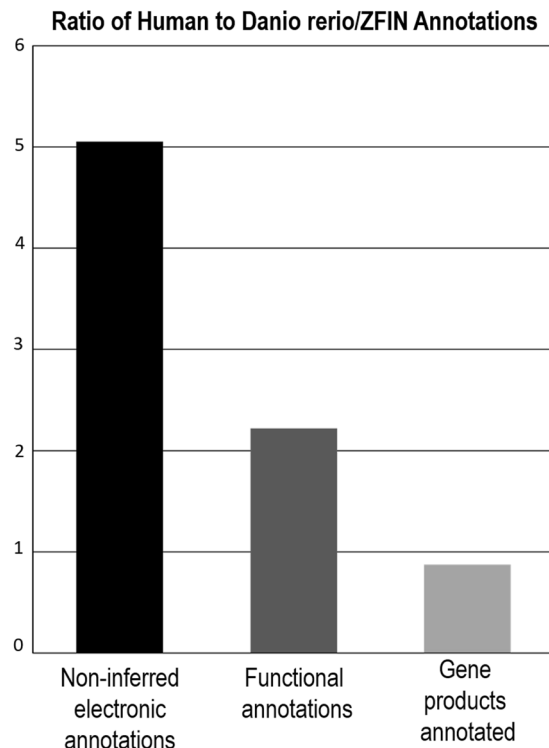


Figure 2. Functional annotations: (A) comparison of zebrafish and human annotations. Zebrafish has larger number of annotated gene products relative to human, 22,171 versus 19,392 (as of April 2017). In human, non-inferred electronic (black bar) and functional annotations (dark grey bar) are 5 times and 2 times better defined respectively than they are in zebrafish. In zebrafish, gene products annotated are slightly better defined than they are in human (light grey bar).

supported these findings (Table 5, DEHP Unique): translation ($q = 2.03E-16$), processing of rRNA ($q = 5.46E-14$) and biogenesis of the ribonucleoprotein complex ($q = 6.20E-11$). Transcripts unique to EE2 were associated with enrichment in the triglyceride metabolic process (Table 5, EE2 Unique, $q = 5.09E-01$), response to dietary excess ($q = 5.47E-01$), and the neutral lipid metabolic process ($q = 7.28E-01$).

In terms of Co-Expression (Supplementary Tables S9,10), in response to DEHP exposure (Table 5, DEHP total; Supplementary Table S10 and Figure S1), we observed an overlap with genes up-regulated with activation of the mTORC1 complex ($q = 6.79E-14$) and genes involved in cholesterol homeostasis ($q = 6.58E-12$).

As the mTORC1 complex is intricately associated with the Insulin-like growth factors (IGF-I and IGF-II) system that regulates metabolism we hypothesized that DEHP exposure could impact liver weight. In order to assess this functional effect we measured liver tissue weights from fish exposed to 100 nM DEHP, 100 nM EE2 or control ethanol exposure. This revealed reduction in the size of the liver in the DEHP exposed fish (Supplementary Figure S1). In response to EE2 exposure (Table 5; Supplementary Table S9), we observed co-expression of genes associated with fatty acid metabolism ($q = 4.65E-04$), as well as genes related to changes in expression observed in hepatocellular carcinoma (HCC) ($q = 3.12E-05$). Analysis of genes unique to exposure to DEHP (Table 5) identified an overlap with genes associated with the sterol regulatory element binding transcription factors (SREBFs) 1 and 2 and the SREBP cleavage-activating protein (SCAP) in the livers of transgenic mice ($q = 4.78E-10$). We found no significant overlap associated with the genes unique to exposure to EE2 (data not shown). Pathway analysis of genes DE expressed by exposure to DEHP revealed five pathways of interest (Supplementary Table S11): FOXA2 and FOXA3 transcription factor networks, metabolic pathways, metabolism of amino acids and derivatives, metabolism of lipids and lipoproteins, and fatty acid, triacylglycerol, and ketone body metabolism. We explored the expression patterns of the genes associated with these pathways in the control and the DEHP-exposed fish (Fig. 4) using heatmaps which show a clear signature associated with DEHP exposure.

Discussion

Environmental chemicals can act through multiple toxicity pathways and induce adverse health outcomes. The relationship between a contaminant and a particular outcome in an individual is dependent on genetic background, target tissue, dose and other factors besides the mechanisms of action (MOA). Transitioning from current risk assessment practices to approaches based on big data collection and integration requires a paradigm shift in how this is executed. A significant challenge to risk assessment is accurately relating chemical impacts across species and stratifying effects or MOAs that are likely to be detrimental to human health. One recent strategy is the adverse outcome (AOP) pathways framework that organizes mechanistic and/or predictive relationships between initial chemical–biological interactions, pathways and networks, and adverse phenotypic outcomes²³.

GO Term	Bonferroni q-value
GO: Biological Process	
DEHP - Total	
organonitrogen compound biosynthetic process	3.87E-16
response to steroid hormone	3.91E-11
cell cycle process	5.23E-11
programmed cell death	2.76E-10
EE2 - Total	
response to endogenous stimulus	7.10E-16
oxoacid metabolic process	1.03E-14
peptide metabolic process	8.23E-12
apoptotic process	1.38E-11
DEHP - Unique	
cell cycle	2.00E-05
RNA processing	1.80E-04
positive regulation of transcription, DNA-templated	7.85E-04
ribonucleoprotein complex biogenesis	1.19E-03
EE2 - Unique	
negative regulation of transcription from RNA polymerase II promoter	2.56E-04
regulation of cell differentiation	1.94E-03
negative regulation of gene expression	1.22E-02
negative regulation of cellular biosynthetic process	1.09E-01
Co-Expression	
DEHP - Total	
Genes whose promoters are bound by MYC [GeneID = 4609], according to MYC Target Gene Database.	1.85E-24
Genes up-regulated in robust Cluster 2 (rC2) of hepatoblastoma samples compared to those in the robust Cluster 1 (rC1).	2.26E-24
Genes up-regulated through activation of mTORC1 complex.	3E-22
EE2 - Total	
Human Liver_Tzur09_1908genes	7.19E-22
Human orthologs of genes up-regulated in the crb ('crash and burn') zebrafish mutant that represents a loss-of-function mutation in BMYB [GeneID = 4605].	6.07E-21
DEHP - Unique	
Genes down-regulated in erythroid progenitor cells from fetal livers of E13.5 embryos with KLF1 [GeneID = 10661] knockout compared to those from the wild type embryos.	4.59E-14
Genes up-regulated through activation of mTORC1 complex.	8.83E-06
EE2 - Unique	
Mouse Liver_White05_638genes	3.21E-06
Rat Liver_Perez-Carreon06_290genes	4.21E-04

Table 2. GO: Biological Process and Co-expression analysis. The top 3,000 ranked DE expressed liver transcripts as determined by microarray analysis from 5.8 nM DEHP and 0.65 nM EE2 exposed adult male zebrafish relative to control fish were mapped to their human homologs using Ensembl homology. GO: Biological Process and co-expression terms and were enriched using ToppFun. The most significant terms for the DEHP and EE2 exposures, and those unique to DEHP (i.e. not present in in EE2 exposed) are presented.

The goal of this study was to assess the effect of environmental levels of DEHP on the liver transcriptome in the adult male zebrafish using a systems level approach. Our analyses utilized a well described whole genome DNA microarray coupled to GOrilla analyses and demonstrated that DEHP deregulates carbohydrate metabolism, including chitin and aminoglycans, and protein synthesis. ToppFun analysis of the significant DE genes from the microarray experiments projected onto their human homologs suggested that DE genes shared by both DEHP/EE2 exposures mapped to pathways related to metabolism, the cell cycle, apoptosis and response to external stimuli. Additionally, DE transcripts unique to DEHP exposure exhibited a signature related to the

Zebra Entrez Gene ID	Probe ID	Gene Symbol	Gene Name	Human Entrez ID	Human Gene Symbol	Log ₂ ratio	FC
FOXA2 and FOXA3 transcription factor networks							
449677	A_15_P117834	cpt1b	carnitine palmitoyltransferase 1B (muscle)	126129	<i>CPT1C</i>	-5.80	-55.72
30262	A_15_P115731	ins	preproinsulin	3630	<i>INS</i>	-3.61	-12.21
317638	A_15_P105778	igfbp1a	insulin-like growth factor binding protein 1a	3484	<i>IGFBP1</i>	-3.54	-11.63
30262	A_15_P110065	ins	preproinsulin	3630	<i>INS</i>	-3.26	-9.58
140815	A_15_P108996	cebpa	CCAAT/enhancer binding protein (C/EBP), alpha	1050	<i>CEBPA</i>	3.20	9.19
573723	A_15_P101211	acadvl	acyl-Coenzyme A dehydrogenase, very long chain	37	<i>ACADVL</i>	-2.51	-5.70
445118	A_15_P107918	g6pca	glucose-6-phosphatase a, catalytic	2538	<i>G6PC</i>	-2.46	-5.50
322493	A_15_P107681	slc2a2	solute carrier family 2 (facilitated glucose transporter), member 2	6514	<i>SLC2A2</i>	2.20	4.59
445118	A_15_P107270	g6pca	glucose-6-phosphatase a, catalytic ^o	2538	<i>G6PC</i>	-2.06	-4.17
325881	A_15_P112672	f2	coagulation factor II (thrombin)	2147	<i>F2</i>	0.10	1.07
Fatty acid, triacylglycerol, and ketone body metabolism							
436636	A_15_P111782	cd36	CD36 antigen	948	<i>CD36</i>	7.88	235.6
768196	A_15_P102657	me1	malic enzyme 1, NADP(+)-dependent, cytosolic	4199	<i>ME1</i>	6.01	64.45
386661	A_15_P112389	scd	stearoyl-CoA desaturase (delta-9-desaturase)	79966	<i>SCD5</i>	5.33	40.22
768196	A_15_P107967	me1	malic enzyme 1, NADP(+)-dependent, cytosolic	4199	<i>ME1</i>	5.23	37.53
393984	A_15_P119395	aacs	acetoacetyl-CoA synthetase	65985	<i>AACS</i>	5.00	32.00
317738	A_15_P108810	elovl6	ELOVL family member 6, elongation of long chain fatty acids (yeast) ^o	79071	<i>ELOVL6</i>	3.68	12.82
317738	A_15_P121489	elovl6	ELOVL family member 6, elongation of long chain fatty acids (yeast)	79071	<i>ELOVL6</i>	3.30	9.85
573723	A_15_P101211	acadvl	acyl-Coenzyme A dehydrogenase, very long chain	37	<i>ACADVL</i>	-2.51	-5.70
393622	A_15_P112797	acsl4a	acyl-CoA synthetase long-chain family member 4a	2182	<i>ACSL4</i>	2.38	5.21
327417	A_15_P109573	hsd17b12a	hydroxysteroid (17-beta) dehydrogenase 12a	51144	<i>HSD17B12</i>	2.20	4.59
Metabolic pathways							
445818	A_15_P112409	cthl	cystathionase (cystathionine gamma-lyase), like ^o	1491	<i>CTH</i>	6.60	97.01
447879	A_15_P114717	zgc:103408	zgc:103408	27231	<i>NMRK2</i>	6.44	86.82
768196	A_15_P102657	me1	malic enzyme 1, NADP(+)-dependent, cytosolic	4199	<i>ME1</i>	6.01	64.45
768196	A_15_P107967	me1	malic enzyme 1, NADP(+)-dependent, cytosolic	4199	<i>ME1</i>	5.23	37.53
436799	A_15_P115180	atp6v1f	ATPase, H ⁺ transporting, V1 subunit F	9296	<i>ATP6V1F</i>	4.47	22.16
393799	A_15_P106842	pycr1	pyrroline-5-carboxylate reductase 1	5831	<i>PYCR1</i>	3.80	13.93
84039	A_15_P108167	bcmo1	beta-carotene 15,15'-monooxygenase 1	53630	<i>BCO1</i>	3.48	11.16
436919	A_15_P111364	ada	adenosine deaminase	100	<i>ADA</i>	3.39	10.48
406651	A_15_P117346	ddc	dopa decarboxylase	1644	<i>DDC</i>	3.31	9.92
406651	A_15_P111461	ddc	dopa decarboxylase	1644	<i>DDC</i>	3.31	9.92
Metabolism of amino acids and derivatives							
445818	A_15_P112409	cthl	cystathionase (cystathionine gamma-lyase), like ^o	1491	<i>CTH</i>	6.60	97.01
393799	A_15_P106842	pycr1	pyrroline-5-carboxylate reductase 1	5831	<i>PYCR1</i>	3.80	13.93
406651	A_15_P117346	ddc	dopa decarboxylase	1644	<i>DDC</i>	3.31	9.92
406651	A_15_P111461	ddc	dopa decarboxylase	1644	<i>DDC</i>	3.31	9.92
572649	A_15_P108171	zgc:112179	zgc:112179	8424	<i>BBOX 1</i>	-2.81	-7.01
114426	A_15_P109191	odc1	ornithine decarboxylase 1	4953	<i>ODC1</i>	2.14	4.41
30665	A_15_P120879	psmb9a	proteasome (prosome, macropain) subunit, beta type, 9a	5698	<i>PSMB9</i>	2.14	4.41
Continued							

Zebra Entrez Gene ID	Probe ID	Gene Symbol	Gene Name	Human Entrez ID	Human Gene Symbol	Log2 ratio	FC
100000775	A_15_P111686	glula	glutamate-ammonia ligase (glutamine synthase) a	2752	GLUL	-2.13	-4.38
321892	A_15_P118448	ckmt1	creatine kinase, mitochondrial 1 ^{''}	1159	CKMT1B	1.93	3.81
399488	A_15_P101243						
	zgc:55813	zgc:55813	6520	SLC3A2	-1.87	-3.66	
Metabolism of lipids and lipoproteins							
436636	A_15_P111782	cd36	CD36 antigen	948	CD36	7.88	235.57
768196	A_15_P102657	me1	malic enzyme 1, NADP(+)-dependent, cytosolic	4199	ME1	6.01	64.45
386661	A_15_P112389	scd	stearoyl-CoA desaturase (delta-9-desaturase)	79966	SCD5	5.33	40.22
768196	A_15_P107967	me1	malic enzyme 1, NADP(+)-dependent, cytosolic	4199	ME1	5.23	37.53
393984	A_15_P119395	aacs	acetoacetyl-CoA synthetase	65985	AACS	5.00	32.00
768298	A_15_P117841	faah2b	fatty acid amide hydrolase 2b	158584	FAAH2	3.83	14.22
317738	A_15_P108810	elovl6	ELOVL family member 6, elongation of long chain fatty acids (yeast) ^{''}	79071	ELOVL6	3.68	12.82
58128	A_15_P109314	fabp7a	fatty acid binding protein 7, brain, a	2173	FABP7	3.56	11.79
58128	A_15_P102880	fabp7a	fatty acid binding protein 7, brain, a	2173	FABP7	3.43	10.78
317738	A_15_P121489	elovl6	ELOVL family member 6, elongation of long chain fatty acids (yeast)	79071	ELOVL6	3.30	9.85

Table 3. DE genes from microarray analysis. in adult male zebrafish exposed to 5.8 nM DEHP relative to controls associated with enriched biological pathways (Metabolic Pathways; Fatty acid, triacylglycerol, and ketone body metabolism; FOXA2 and FOXA3 transcription factor networks; Metabolism of amino acids and derivatives; Metabolism of Lipids and Lipoproteins). Genes with the greatest fold change difference in DEHP exposed relative to control are presented. An expanded list of DE genes is presented in Supplementary Table S6.

regulation of transcription and translation. Co-expression analysis revealed overlap with gene expression patterns observed in apical endpoints such as liver cancer, and the up-regulation of genes in the mTORC1 pathway. An independent experiment exploited high throughput RNA-sequencing and this data supported these results, highlighting that DEHP exposure affected metabolic processes, particularly lipid metabolism. The DE mRNA signature unique to DEHP exposure was associated with GO terms related to gene expression, protein synthesis, lipid/fatty acid metabolism and RNA metabolism. Co-expression analysis again revealed similarity to genes linked with adverse liver disease outcomes. Finally, we compared microarray and RNA-Seq analyses, based on the projection of zebrafish genes onto their human orthologs and subsequent analysis using the ToppGene Suite for gene list enrichment analysis and the following pathway annotations; BioSystems: Pathway Interaction Database, BioSystems: REACTOME and BioSystems: KEGG. This comparison uncovered five pathways: 'FOXA2 and FOXA3 transcription factor networks,' 'Metabolic pathways,' 'metabolism of amino acids and derivatives,' 'metabolism of lipids and lipoproteins,' and 'fatty acid, triacylglycerol, and ketone body metabolism' shared between the microarray and sequencing experiments.

The Benefits of Using Human Annotations in a Zebrafish Study. Presently the zebrafish genome is not as well characterized as the human genome and the level of functional annotation lags behind. However 70% of protein-coding human genes are related to genes found in the zebrafish and 84% of genes known to be associated with human disease have a zebrafish counterpart³⁴. It is therefore worthwhile to consider the human orthologs of the zebrafish genes in systems level analyses. As shown in Fig. 2⁶ in terms of the ratio of human to zebrafish annotations, there are 5 times more non-inferred electronic and 2.2 times more functional annotations in the human genome relative to the zebrafish (based on the Gene Ontology Consortium data April 2017). Utilizing GO terms for human orthologs provides a deeper systems level analysis than that obtained with zebrafish genes. We achieved a greater number of enriched terms in mRNA signatures unique to EE2 using human orthologs: enrichment analysis of the zebrafish genes in GOrilla uncovered no significant results, whereas analysis of the corresponding human orthologs with ToppFun identified significant enriched terms. A limitation to this approach is the absence of specific genes, for example vitellogenins (VTGs) which do not have orthologs in humans²¹. However for a comparative analysis between zebrafish and humans, however, these limitations are outweighed by the benefits of improved annotations and the comprehensive systems analysis afforded.

Exposure to both DEHP and EE2 alter metabolic processes in the liver. Our data suggest that exposure to both EE2 and DEHP leads to the differential expression of genes involved in metabolic pathways representing molecular initiating events (MIEs). Since the liver is the primary site of metabolism within the body, the enrichment of metabolic pathways in both the DEHP and EE2 exposures was not unexpected. Among these common enrichment terms were changes to organic acid metabolism, in particular carboxylic acid metabolism³⁵.

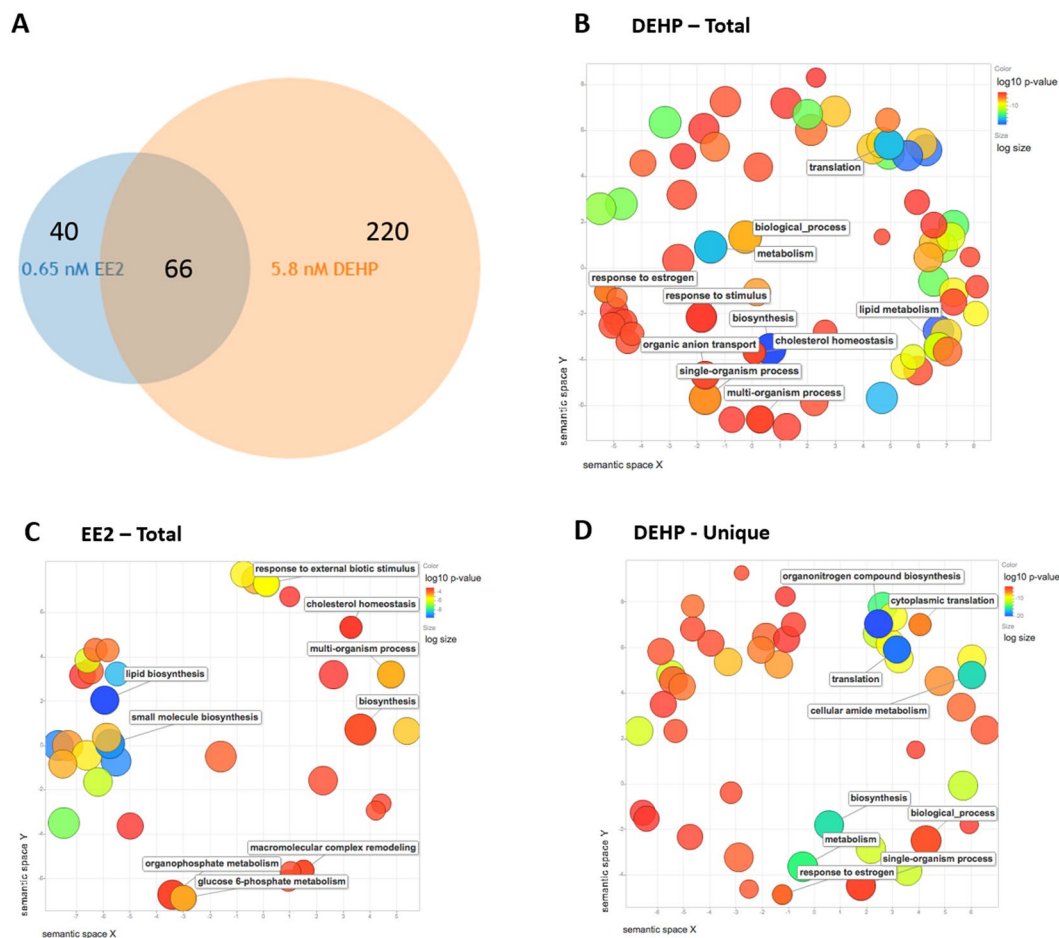


Figure 3. Functional Analyses RNA-Seq data. **(A)** Overlap of the significant DE expressed liver transcripts (FDR < 0.4) from 5.8 nM DEHP and 0.65 nM EE2 exposed adult male zebrafish relative to control fish as determined by DESeq. 2. **(B–D)** Gene Ontology Biological Process analyses: Scatterplots shows the cluster representatives (i.e. terms remaining after the redundancy reduction) in a two dimensional space derived by applying multidimensional scaling to a matrix of the GO terms' semantic similarities. Bubble color indicates the p-value (legend in upper right-hand corner); size indicates the frequency of the GO term in the underlying GOA database (bubbles of more genes are larger). GO BP analysis of DE genes in **(B)** DEHP and **(C)** EE2 exposed livers. **(D)** GO BP analysis of DE genes unique to DEHP (not DE in EE2 exposed).

Co-expression results from ToppFun analysis identified overlaps between the DE mRNAs and the human orthologs of genes upregulated in the zebrafish *crb* (“crash and burn”) mutant, this mutation induces a loss-of-function of the gene *bmyb*, a transcription factor³⁶. Loss of *bmyb* function can cause genome instability, and adult *crb* zebrafish heterozygotes have an increased cancer susceptibility³⁶. Our results show that exposure to both DEHP and EE2 can mimic the effects of *bmyb* inactivation, suggesting a possible role for both molecules in carcinogenesis.

Unique effects of DEHP exposure indicate links with non-alcoholic fatty liver disease (NAFLD).

The list of significant DE mRNAs overlaps with genes up-regulated in the liver of transgenic mice overexpressing sterol regulatory element binding transcription factors 1 and 2 (*SREBF1* and *SREBF2*), and with low levels of SREBP cleavage-activating protein (*SCAP*) chaperone³⁷. SREBF facilitates the development of NAFLD by increasing the synthesis of fatty acids within hepatocytes³⁸. *SCAP* is required for the activation of SREBPs, and knockout of *SCAP* in mice has been shown to reduce the rate of fatty acid synthesis in the liver^{39,40}. Another xenoestrogen bisphenol A (BPA) was shown recently to produce hepatosteatosis in zebrafish and human hepatocytes by up-regulating the endocannabinoid system. Hepatosteatosis, was associated with an increase in the liver levels of the obesogenic endocannabinoids 2-arachidonoylglycerol and anandamide and a concomitant decrease in palmitoylethanolamide⁴¹. Furthermore, we recently demonstrated that chronic BPA exposure impacts the hepatic epigenome in adult male zebrafish with altered gene expression signatures associated with non-alcoholic fatty liver disease (NAFLD) and cell cycle⁴². Similar results have been obtained in our laboratory with di-isononyl phthalate (DiNP) which upregulates orexigenic signals and causes hepatosteatosis together with deregulation of the peripheral endocannabinoid system (ECS) and lipid metabolism⁴³.

GO Term	Bonferroni q-value
DEHP Total	
translation	6.09E-14
lipid metabolic process	1.05E-08
lipid biosynthetic process	7.01E-08
response to estrogen	2.94E-04
cytoplasmic translation	5.19E-04
response to hypoxia	1.24E-02
EE2 - Total	
lipid biosynthetic process	8.58E-07
lipid metabolic process	2.58E-04
triglyceride metabolic process	2.05E-02
triglyceride-rich lipoprotein particle remodeling	6.13E-02
triglyceride catabolic process	9.03E-02
regulation of insulin-like growth factor receptor signaling pathway	1.03E-01
DEHP - Unique	
translation	3.10E-16
fatty acid biosynthetic process	1.26E-15
protein metabolic process	1.29E-02

Table 4. RNA-Seq: GOrilla functional enrichment analysis. Significant DE expressed liver transcripts (FDR < 0.4) as determined by RNA-Seq from 5.8 nM DEHP and 0.65 nM EE2 exposed adult male zebrafish relative to control fish. The most significant terms for the DEHP and EE2 exposures, and those unique to DEHP (i.e. not present in in EE2 exposed) are presented.

Non-estrogenic effects of DEHP relate to changes in translation and membrane ruffling. When we analyzed DEHP's global effects on the liver transcriptome, in parallel we assessed the effects of EE2. This allowed us to identify the estrogenic effects of DEHP, as well as separate non-estrogenic effects by considering DE mRNAs unique to DEHP exposure and therefore not regulated by EE2. Analysis of these unique mRNAs identified the non-estrogenic effects of DEHP exposure. GO analysis of zebrafish terms indicated that DEHP uniquely induced changes in translational initiation, a signature not observed with EE2 exposure. This is supported by data from both the microarray and RNA-Seq experiments.

Co-expression analysis of genes unique to DEHP exposure supported the connection between translational changes and DEHP exposure. We observed that genes DE in response to DEHP exposure overlapped with genes up-regulated in relation to the mTORC1 pathway; mTORC1 regulates protein synthesis by accepting growth factors to activate translation, consequently making it a major effector of cell growth and proliferation^{44,45}. Dysregulation of mTORC1 is associated with the development of an increasing number of pathologies, including cancer⁴⁶, obesity^{47,48}, type 2 diabetes^{49,50}, and NAFLD^{51,52}. Along with Akt and S6K1, mTORC1 is a key component of the insulin signaling pathway, and has been identified as an enhancer of SREBP1c's role as master transcription factor in lipid synthesis⁵³; mTORC1 regulates SREBP by controlling the nuclear entry of a phosphatidic acid phosphatase called lipin 1, which mediates the effects of mTORC1 on SREBP target gene, SREBP promoter activity, and nuclear SREBP protein abundance⁵². Intriguingly, hyperactivity of mTORC1 in liver cells has been shown to protect against steatosis in mice⁵¹; deletion of *Tsc1* in mouse liver cells, which codes hamartin (TSC1) of the tuberous sclerosis complex (TSC) and controls the activity of mTORC1⁵⁴, specifically protected cells from high-fat diet induced, Akt-mediated steatosis through restriction of S6K1 independent of Akt suppression⁵¹.

We examined the most significantly deregulated mRNAs that are part of the mTORC1 pathway. *PPP1R15A*, *LDLR*, *PNP*, *IFRD1*, *NUPR1*, *SSR1* and *PNO1* were all downregulated in response to DEHP exposure. Protein Phosphatase 1 Regulatory Subunit (*PPP1R15A*) mRNA levels are increased following stressful growth arrest conditions⁵⁵. The low density lipoprotein receptor (*LDLR*) gene encodes a cell surface protein involved in receptor-mediated endocytosis of specific ligands⁵⁶. Purine Nucleoside Phosphorylase (*PNP*) encodes an enzyme which reversibly catalyzes the phosphorolysis of purine nucleosides⁵⁷. Interferon Related Developmental Regulator 1 (*IFRD1*) is an immediate early gene that encodes a protein related to interferon-gamma, it can function as a transcriptional co-activator/repressor and control the growth and differentiation of specific cell types during embryonic development and tissue regeneration⁵⁸. Nuclear Protein 1 Transcriptional Regulator (*NUPR1*) encodes a chromatin-binding protein that converts stress signals into a program of gene expression that empowers cells with resistance to the stress induced by a change in their microenvironment⁵⁹. The signal sequence receptor 1 (*SSR1*) is a glycosylated endoplasmic reticulum (ER) membrane receptor associated with protein translocation across the ER membrane⁶⁰. *ELOVL6*, *HMGCR*, *DHCR7*, *CYP51A1*, *HMGCS1*, *ME1*, *SQLE*, *SC5D*, *CTH*, *BCAT1*, *ACLY*, *ELOVL5*, *ACACA*, *HSPA5*, *HSPD1*, *IDH1* and *ELOVL6* were all upregulated in response to DEHP exposure. Fatty Acid Elongase 6 (*ELOVL6*) uses malonyl-CoA as a 2-carbon donor in the first and rate-limiting step of fatty acid elongation⁶¹. 3-Hydroxy-3-Methylglutaryl-CoA Reductase (*HMGCR*) is the rate-limiting enzyme for cholesterol synthesis. 7-Dehydrocholesterol Reductase (*DHCR7*) encodes an enzyme that removes the C(7-8) double bond in the B ring of sterols and catalyzes the conversion of 7-dehydrocholesterol to cholesterol. Cytochrome P450 Family 51 Subfamily A Member 1 (*CYP51A1*) encodes a member of the cytochrome

GO Term	Bonferroni q-value
GO: Biological Process	
DEHP - Total	
cotranslational protein targeting to membrane	1.01E-19
nuclear-transcribed mRNA catabolic process, nonsense-mediated decay	1.79E-16
translational initiation	2.84E-15
peptide biosynthetic process	6.43E-14
EE2 - Total	
small molecule biosynthetic process	2.27E-10
lipid biosynthetic process	3.87E-09
fatty acid metabolic process	7.42E-08
cholesterol metabolic process	9.28E-06
DEHP - Unique	
protein targeting to ER	2.86E-19
translation	2.03E-16
rRNA processing	5.46E-14
ribonucleoprotein complex biogenesis	6.20E-11
EE2 - Unique	
triglyceride metabolic process	5.09E-01
purine nucleoside monophosphate metabolic process	5.22E-01
response to dietary excess	5.47E-01
neutral lipid metabolic process	7.28E-01
Co-Expression	
DEHP - Total	
Genes up-regulated through activation of mTORC1 complex.	6.79E-14
Genes involved in cholesterol homeostasis.	6.58E-12
EE2 - Total	
Genes down-regulated in hepatocellular carcinoma (HCC) compared to normal liver samples.	3.12E-05
Genes encoding proteins involved in metabolism of fatty acids.	4.65E-04
DEHP - Unique	
Molecular timetable composed of 162 time-indicating genes (182 probes) in the peripheral (liver) clock.	2.49E-12
Genes up-regulated in liver from mice transgenic for SREBF1 or SREBF2 [GeneID = 6720, 6721] and down-regulated in mice lacking SCAP [GeneID = 22937].	4.78E-10

Table 5. RNA-Seq: GO: Biological Process and Co-expression analysis. Significant DE expressed liver transcripts (FDR < 0.4) as determined by RNA-Seq from 5.8 nM DEHP and 0.65 nM EE2 exposed adult male zebrafish relative to control fish were mapped to their human homologs using Ensembl homology. GO: Biological Process and co-expression terms and were enriched using ToppFun. The most significant terms for the DEHP and EE2 exposures, and those unique to DEHP (i.e. not present in in EE2 exposed) are presented. Expanded lists are found in Supplementary Tables S7–S10.

P450 superfamily and participates in the synthesis of cholesterol by catalyzing the removal of the 14 α -methyl group from lanosterol. 3-Hydroxy-3-Methylglutaryl-CoA Synthase 1 (*HMGCS1*) is an enzyme that condenses acetyl-CoA with acetoacetyl-CoA to form HMG-CoA, which is the substrate for HMG-CoA reductase. Malic Enzyme 1 (*ME1*) encodes a cytosolic, NADP-dependent enzyme that generates NADPH for fatty acid biosynthesis. The activity of this enzyme, the reversible oxidative decarboxylation of malate, links the glycolytic and citric acid cycles. Among its related pathways are Metabolism and Regulation of lipid metabolism by Peroxisome proliferator-activated receptor alpha (PPAR α). Squalene Epoxidase (*SQLE*) catalyzes the first oxygenation step in sterol biosynthesis. Sterol-C5-Desaturase (*SC5D*) encodes an enzyme of cholesterol biosynthesis. Branched Chain Amino Acid Transaminase 1 (*BCAT1*) encodes the cytosolic form of the enzyme branched-chain amino acid transaminase. ATP Citrate Lyase (*ACLY*) is the primary enzyme responsible for the synthesis of cytosolic acetyl-CoA in many tissues. ELOVL Fatty Acid Elongase (*ELOVL5*) encodes a multi-pass membrane protein that is localized in the endoplasmic reticulum and is involved in the elongation of long-chain polyunsaturated fatty acids. Acetyl-CoA Carboxylase Alpha (*ACACA*) is regulated by the phosphorylation/dephosphorylation of targeted serine residues and by allosteric transformation by citrate or palmitoyl-CoA. Heat Shock Protein Family A (*Hsp70*) Member (HSPA5) is localized in the lumen of the endoplasmic reticulum (ER), and is involved in the folding and assembly of proteins in the ER⁵². As this protein interacts with many ER proteins, it may play a key role in monitoring protein transport through the cell. Heat Shock Protein Family D (*Hsp60*) Member 1 (HSPD1) encodes a member of the chaperonin family.

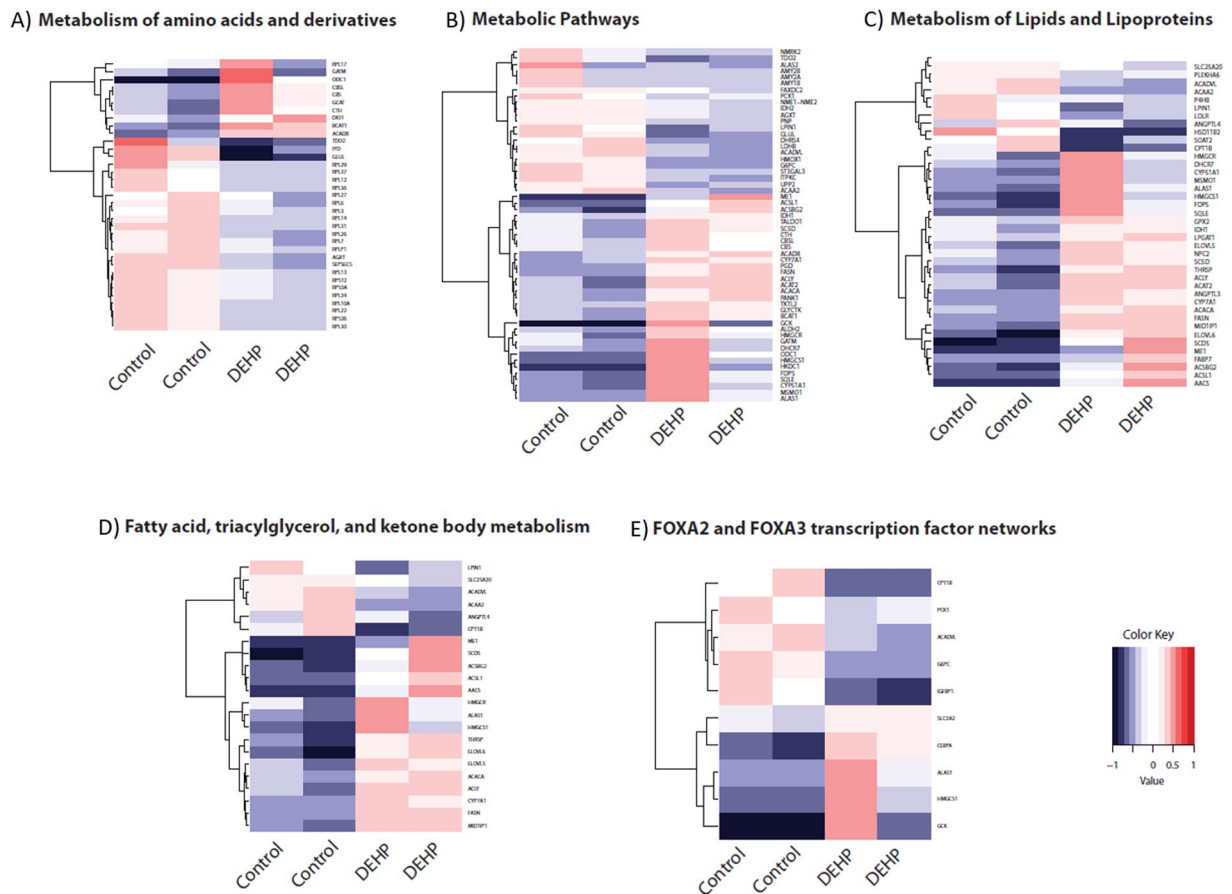


Figure 4. Heatmaps of significantly DE liver mRNAs as determined by DESeq. 2. (FDR < 0.4) in adult male zebrafish exposed to 5.8 nM DEHP relative to controls. DE genes associated with enriched biological pathways (A) Metabolism of amino acids and derivatives, (B) Metabolic Pathways, (C) Metabolism of Lipids and Lipoproteins, (D) Fatty acid, triacylglycerol, and ketone body metabolism and (E) FOXA2 and FOXA3 transcription factor networks are plotted. Red and blue boxes indicate relative over- and under-expression with respect to a reference which is calculated as the mid-point between the control and exposed groups.

We observed a reduction in the size of the liver in adult zebrafish in response to a seven day 100 nM DEHP exposure (Supplementary Figure S1). The mTOR signaling pathway senses and integrates a variety of environmental cues to regulate organismal growth and homeostasis⁴⁵. DEHP exposure in the nematode *Caenorhabditis elegans* was found to be acutely toxic, with decreases in body length and egg number per worm observed after 24 h of DEHP exposure⁶³. DEHP has also been shown to decrease serum estradiol levels and aromatase expression, prolong estrous cycles, and cause anovulation in animal and culture models and directly. It inhibits antral follicle growth via a mechanism that partially includes reduction in levels of estradiol production and decreased expression of cell cycle regulators⁶⁴.

Other significant GO terms unique to DEHP exposure included ruffle assembly and organization - the formation of actin rich membrane protrusions typically used for cell motility⁶⁵. In the context of the liver, ruffles form in response to insulin signaling. Mice with the fatty liver dystrophy (*fld*) mutation lose the ability to properly form membrane ruffles, the result of impairments in insulin-mediated cytoskeleton dynamics⁶⁶. One of the hallmarks of this mutation is the formation of a fatty liver, similar to what is observed in NAFLD⁶⁷. Furthermore, in muscle cells, defective actin remodeling is associated with the development of insulin resistance⁶⁸; defined as the inability of cells to be regulated by insulin signaling, insulin resistance is associated with adverse outcomes such as type 2 diabetes, cardiovascular disease, and chronic liver disease^{69–71}. Recently, exposure to DEHP has been implicated with disruption of the insulin signaling pathway of both rats and the human cell line L-02 by activating PPARG, reducing the liver's ability to maintain glucose homeostasis and contributing to insulin resistance⁴. The significant enrichment of an insulin-mediated pathway – ruffle organization – supports the role of DEHP as a disrupter of insulin function, including the potential to inhibit insulin signaling in liver cells. Analysis of the signatures unique to DEHP exposure has identified changes related to translation and membrane ruffling; combined with our co-expression analysis, these perturbations support an association between DEHP exposure and NAFLD.

Comparison of microarray and RNA-Seq platforms. Both the microarray and RNA-Seq analyses of the transcriptome identified perturbations in gene expression. While RNA-Seq is a newer more reproducible technology with less background noise, and a greater dynamic range to measure gene expression, DNA microarrays are

still commonly utilized due to their cost and ease³⁰. In the context of our results, we were interested in comparing microarray and sequencing analyses from the two different experiments where zebrafish were exposed to 5.8 nM DEHP. In general, we found that the results of our sequencing analysis supported those of our microarray analysis; many of the same GO and co-expression terms were enriched in both analyses, including genes up-regulated with the loss of *bmyb* and changes in translation initiation in response to exposure to DEHP.

In addition, the sequencing analysis uncovered GO terms that strengthen the connection between DEHP exposure and the development of adverse outcomes in the liver. For example, analysis of the microarray data derived signatures unique to DEHP exposure identified enrichments in membrane ruffling, a process regulated by insulin in liver cells⁷². Analysis of the RNA-Seq data determined that exposure to DEHP enriched the same set of genes associated with changes in regulation of *SREBF1* and *SREBF2*. These transcription factors are significant in the development of liver pathologies and insulin resistance, and support previous studies associating exposure to DEHP with dysregulation of the insulin pathway⁴. Taken together, the combined genomic platforms identified molecular initiating and key events in the liver transcriptome that are conducive to the development of insulin resistance.

Bone morphogenic protein 2 (BMP2) and the epigenetic effects of DEHP exposure. Analysis of DEHP's unique mRNA signature obtained from the microarray experiment identified an overlap with genes that were down-regulated in the uterus upon knockdown of bone morphogenic protein 2 (BMP2) (Supplementary Table S3). In the uterus, BMP signaling is necessary for embryonic development, and BMP2 in particular is critical for embryonic implantation by inducing decidualization⁷³, a rapid remodeling of uterine endometrial stroma into epithelial decidual cells that is critical for the progress of implantation⁷⁴. Knockout of *BMP2* in mice renders mothers infertile due to lack of signaling to the stromal cells to begin decidualization⁷³. Outside of the uterus, exposure to DEHP has been associated with decreased expression of BMPs, including BMP2, in dam mesenchymal stem cells⁷⁵. DEHP mediates its adverse effects by interfering with signaling mechanisms involved in oocyte growth (VTG), maturation via the bone morphogenetic protein-15 (BMP15), luteinizing hormone receptor (LHR), membrane progesterone receptors (mPRs) and ovulation (cyclooxygenase (COX)-2 (pts2)), thereby deeply impairing ovarian functions⁷⁶; female zebrafish exposed to environmentally relevant doses of DEHP exhibited a significant decrease in fecundity with diminished rates of ovulation and embryo production.

DEHP causes epigenetic effects impacting gene expression in the developing testis⁷⁷. Alterations in DNA methylation patterns caused by maternal exposure have been indicated to play a key role in DEHP-mediated testicular toxicity. Maternal exposure to DEHP at 500 mg/kg/d induces testicular dysgenesis syndrome in fetuses and embryos⁷⁸. DEHP exposure has also been shown to significantly inhibit gap-junctional intercellular communication, likely providing the initial stimulus that enables cell transformation, and facilitates development of these cells into tumors⁷⁹. *In utero* DEHP exposure delays maturation of fetal Leydig cells, with reduced mineralocorticoid receptor (MR; NR3C2) expression in the adult Leydig cell⁸⁰. Activation of MR induces androgen synthesis⁸¹, whereas its inhibition blocks testosterone synthesis in the Leydig cell⁸². Interestingly, it has been shown that DNA methylation and histone modification play a role in the regulation of *BMP2*^{83,84}. Taken together, these studies suggest that DEHP could regulate the expression of *BMP2* via epigenetic modification.

Conclusion

The focus of this study was to examine the global effects of exposure to environmental levels of the plasticizer DEHP in the adult male zebrafish and uncover perturbations that represent MIEs that can potentially lead to adverse outcomes. Our results indicate that exposure to DEHP and EE2 leads to the expression of liver genes related to the mTORC1 complex, fatty acid metabolism and the development of non-alcoholic fatty liver disease. Analysis of only the DE genes associated with exposure to DEHP reveals potentially non-estrogenic perturbations in eukaryotic translation and insulin signaling after exposure. This data supports previous studies that implicate DEHP as increasing the risk of developing NAFLD.

Materials and Methods

Ethinylestradiol and (2-ethylhexyl) phthalate exposure studies in zebrafish. For both the microarray and RNA-Seq experiments adult male zebrafish (*Danio rerio*) were housed in aquaria that were individually heated using a 100 W aquarium heater to maintain a temperature of 26–29 °C, and the light–dark cycle was 14:10 h. The water pH ranged from 7.0 to 7.6 throughout the duration of the experiment. Aeration and filtration were provided using sponge filters. Fish were fed two times a day with commercial flaked fish food (Tetra, Germany). Fish were acclimated for one week prior to commencing the experiments. Three tanks (80 l/tank) with 40 animals each were prepared for the different experimental groups, one containing water with 5.8 nM of (2-ethylhexyl) phthalate (DEHP), one tank containing water with 0.65 nM of ethinylestradiol 17- α (EE2), and one tank containing water and ethanol as negative exposure control. The levels of DEHP are considered “environmentally relevant¹⁹,” and previous studies have determined that exposing model organisms to these concentrations is associated with altered effects on the transcriptome^{21,76,85,86}. All chemicals were dissolved in ethanol and stock working solutions were prepared, from which the working experimental concentrations were prepared. The nominal exposures utilized a continuous flow-through system. Following a 21 day exposure the fish were sacrificed and sampled for liver tissue. Tissue samples were immediately frozen in liquid nitrogen and stored at –70 °C. For the liver mass experiments, animals were exposed to 100 nM of DEHP or EE2 for seven days, and the animals were sampled for liver tissue. All procedures involving zebrafish were performed in accordance with The University of California San Diego, Institutional Animal Care and Use Committee (IACUC) guidelines. All the experiments described above were approved by the IACUC and were performed in accordance with institutional guidelines and regulations.

RNA Extraction. Isolation of total RNA from zebrafish liver samples was performed using TRIzol reagent (Invitrogen) and the extracted RNA were further purified using the RNeasy Mini kit (Qiagen, Valencia, CA). All RNA was subjected to on-column digestion of DNA during RNA purification from cells, to ensure highly pure RNA free from DNA contamination. The concentrations were determined by absorbance readings (OD) at 260 nm using an ND-1000 (Nanodrop, Wilmington, DE). RNA was further assessed for integrity with the 6000 Nano LabChip assay from Agilent, (Santa Clara, CA). Only RNA samples with a RIN score of >7.0 were used for genomic analyses. There were 18 samples in total (six EE2 exposed, six DEHP exposed, six control fish), selected for RNA extraction.

Microarray Analysis. Using the Low RNA Input Fluorescent Linear Amplification Kit (Agilent), 100 ng of total RNA were converted into fluorescently labeled Cy3 cRNA. Unincorporated nucleotides of fluorescent targets were removed using RNeasy (Qiagen). Absorbance (OD) at 260 nm was used to quantify cRNA concentrations, and absorbance at 550 nm was used to measure Cy3 dye incorporation. Microarray hybridization was only carried out with cRNA that had an incorporation efficiency of 9 pmol/μg or greater.

We utilized the Agilent *D. Rerio* Oligo Microarray 4 × 44 K G2519F (015064), array design A-MEXP-1396 (Santa Clara, CA). cRNA target hybridization to the zebrafish microarray was carried out in accordance with single color Agilent Hybridization protocols, and have been described previously²⁵. Array data were collected using an Agilent Microarray Scanner and Feature Extraction Software (v10.5), and deposited in the ArrayExpress Database, accession number E-TABM-547. Though Agilent's Feature Extraction Software (v10.5) provides high quality expression reports for microarrays, the data still needed to be normalized to remove background technical noise and other subtle biases caused by hybridization. For this experiment, statistical analysis of the microarray experiment involved two steps: normalization and sorting of genes according to interest. All samples were normalized simultaneously using the multiple-loess technique⁸⁷. Log ratios for each probe (technical replicates) were calculated separately, then averaged over the biological replicate microarrays. The data was sorted using the interest statistic, which reflects the understanding that the gene with a greater absolute fold change is potentially more interesting, (which we have described in greater detail previously)^{24–26}. The design of the interest statistic was based on ideas borrowed from the software package Focus⁸⁸.

RNA sequencing (RNA-Seq). To prepare RNA-Seq libraries using the TruSeq RNA Sample Prep Kit (Illumina, San Diego, CA), 100–200 ng of total RNA was used following the protocol described by the manufacturer. High throughput sequencing (HTS) was performed using an Illumina GAIIX with each sample sequenced to a minimum depth of ~2 million reads. A single end 50 cycle sequencing strategy was employed. Data were subjected to Illumina quality control (QC) procedures (>80% of the data yielded a Phred score of 30). RNA-Seq data has been submitted to the NCBI Gene Expression Omnibus, accession number GSE100367. Secondary analysis was carried out on an OnRamp Bioinformatics Genomics Research Platform (OnRamp Bioinformatics, San Diego, CA)³⁰. OnRamp's advanced Genomics Analysis Engine utilized an automated RNA-seq workflow to process the data^{89,90}, including (1) data validation and quality control, (2) read alignment to the zebrafish genome (GRCz10) using TopHat2⁹¹, which revealed >73% mapping, (3) generation of gene-level count data with HTSeq⁹², and (4) differential expression analysis with DESeq 2⁹³. Transcript count data from DESeq 2 analysis of the samples were sorted according to their adjusted p-value (or q-value), which is the smallest false discovery rate (FDR) at which a transcript is called significant. FDR is the expected fraction of false positive tests among significant tests and was calculated using the Benjamini-Hochberg multiple testing adjustment procedure^{30,31,93,94}.

Gene Ontology and Pathway Analyses. The Gene Ontology Enrichment Analysis and Visualization Tool (GORilla) was used to search enriched GO terms associated with DEHP and EE2 exposures^{27,95}. Data was further analyzed with the GO summarization tool REVIGO⁹⁶ which combines redundant terms into a single, representative term based on a simple clustering algorithm relying on semantic similarity measures. We exploited Ensembl homology to append a human gene ID to a given zebrafish gene ID, in order to permit systems analysis using the 'Transcriptome, ontology, phenotype, proteome, and pharmacome annotations based gene list functional enrichment analysis' (Toppfun) tool and the richer Gene Ontology content available for human compared to zebrafish⁶. ToppFun sources content from multiple databases, including KEGG, WikiPathways, and REACTOME^{97–99}. To ensure that only the most relevant terms were selected, we applied Bonferroni FDR correction.

References

1. Wang, Y. & Zhou, J. Endocrine disrupting chemicals in aquatic environments: A potential reason for organism extinction? *Aquatic Ecosystem Health & Management* **16**, 88–93, <https://doi.org/10.1080/14634988.2013.759073> (2013).
2. Franken, C. *et al.* Phthalate-induced oxidative stress and association with asthma-related airway inflammation in adolescents. *Int J Hyg Environ Health*. <https://doi.org/10.1016/j.ijheh.2017.01.006> (2017).
3. Diamanti-Kandarakis, E. *et al.* Endocrine-disrupting chemicals: an Endocrine Society scientific statement. *Endocr Rev* **30**, 293–342, <https://doi.org/10.1210/er.2009-0002> (2009).
4. Zhang, W. *et al.* Di-(2-ethylhexyl) phthalate could disrupt the insulin signaling pathway in liver of SD rats and L02 cells via PPARgamma. *Toxicology and applied pharmacology* **316**, 17–26, <https://doi.org/10.1016/j.taap.2016.12.010> (2017).
5. Li, L. *et al.* The molecular mechanism of bisphenol A (BPA) as an endocrine disruptor by interacting with nuclear receptors: insights from molecular dynamics (MD) simulations. *PLoS One* **10**, e0120330, <https://doi.org/10.1371/journal.pone.0120330> (2015).
6. Baker, M. E. & Hardiman, G. Transcriptional analysis of endocrine disruption using zebrafish and massively parallel sequencing. *J Mol Endocrinol* **52**, R241–256, <https://doi.org/10.1530/JME-13-0219> (2014).
7. Murnyak, G. *et al.* Emerging contaminants: presentations at the 2009 Toxicology and Risk Assessment Conference. *Toxicology and applied pharmacology* **254**, 167–169, <https://doi.org/10.1016/j.taap.2010.10.021> (2011).
8. Paterni, I., Granchi, C. & Minutolo, F. Risks and Benefits Related to Alimentary Exposure to Xenoestrogens. *Crit Rev Food Sci Nutr* **0**, <https://doi.org/10.1080/10408398.2015.1126547> (2016).

9. Gassman, N. R., Coskun, E., Jaruga, P., Dizdaroglu, M. & Wilson, S. H. Combined Effects of High-Dose Bisphenol A and Oxidizing Agent (KBrO) on Cellular Microenvironment, Gene Expression, and Chromatin Structure of Ku70-deficient Mouse Embryonic Fibroblasts. *Environ Health Perspect*. <https://doi.org/10.1289/ehp237> (2016).
10. Huang, B. *et al.* Bisphenol A Represses Dopaminergic Neuron Differentiation from Human Embryonic Stem Cells through Downregulating the Expression of Insulin-like Growth Factor 1. *Mol Neurobiol*. <https://doi.org/10.1007/s12035-016-9898-y> (2016).
11. Snyder, R. W. *et al.* Metabolism and disposition of bisphenol A in female rats. *Toxicology and applied pharmacology* **168**, 225–234, <https://doi.org/10.1006/taap.2000.9051> (2000).
12. Yang, O., Kim, H. L., Weon, J. I. & Seo, Y. R. Endocrine-disrupting Chemicals: Review of Toxicological Mechanisms Using Molecular Pathway. *Analysis*. **20**, 12–24, <https://doi.org/10.15430/jcp.2015.20.1.12> (2015).
13. Friis, R. Vol. 1 (ABC-CLIO, 2012).
14. Pak, V. M., Nailon, R. E. & McCauley, L. A. Controversy: neonatal exposure to plasticizers in the NICU. *MCN Am J Matern Child Nurs* **32**, 244–249, <https://doi.org/10.1097/01.NMC.0000281965.45905.c0> (2007).
15. Rudel, R. A. *et al.* Identification of Selected Hormonally Active Agents and Animal Mammary Carcinogens in Commercial and Residential Air and Dust Samples. *Journal of the Air & Waste Management Association* **51**, 499–513, <https://doi.org/10.1080/10473289.2001.10464292> (2001).
16. Bernard, L., Cuff, R., Breyse, C., Decaudin, B. & Sautou, V. Migrability of PVC plasticizers from medical devices into a simulant of infused solutions. *Int J Pharm* **485**, 341–347, <https://doi.org/10.1016/j.ijpharm.2015.03.030> (2015).
17. Chen, H. *et al.* Di(2-ethylhexyl) phthalate exacerbates non-alcoholic fatty liver in rats and its potential mechanisms. *Environ Toxicol Pharmacol* **42**, 38–44, <https://doi.org/10.1016/j.etap.2015.12.016> (2016).
18. Zhang, W. *et al.* The effects of di 2-ethyl hexyl phthalate (DEHP) on cellular lipid accumulation in HepG2 cells and its potential mechanisms in the molecular level. *Toxicol Mech Methods* **1–8**, <https://doi.org/10.1080/15376516.2016.1273427> (2017).
19. Clark, K., Cousins, I. T. & Mackay, D. In *Series Anthropogenic Compounds: Phthalate Esters* (ed. Charles A. S.) 227–262 (Springer Berlin Heidelberg, 2003).
20. Jackson, J. & Sutton, R. Sources of endocrine-disrupting chemicals in urban wastewater, Oakland, CA. *Sci Total Environ* **405**, 153–160, <https://doi.org/10.1016/j.scitotenv.2008.06.033> (2008).
21. Maradonna, F. *et al.* Assay of vtg, ERs and PPARs as endpoint for the rapid *in vitro* screening of the harmful effect of Di-(2-ethylhexyl)-phthalate (DEHP) and phthalic acid (PA) in zebrafish primary hepatocyte cultures. *Toxicol In Vitro* **27**, 84–91, <https://doi.org/10.1016/j.tiv.2012.09.018> (2013).
22. Chikae, M. *et al.* Effects of bis(2-ethylhexyl) phthalate and benzo[a]pyrene on the embryos of Japanese medaka (*Oryzias latipes*). *Environ Toxicol Pharmacol* **16**, 141–145, <https://doi.org/10.1016/j.etap.2003.11.007> (2004).
23. Garcia-Reyero, N. Are adverse outcome pathways here to stay? *Environ Sci Technol* **49**, 3–9, <https://doi.org/10.1021/es504976d> (2015).
24. Ogawa, S. *et al.* A nuclear receptor corepressor transcriptional checkpoint controlling activator protein 1-dependent gene networks required for macrophage activation. *Proc Natl Acad Sci USA* **101**, 14461–14466, <https://doi.org/10.1073/pnas.0405786101> (2004).
25. Baker, M. E. *et al.* Analysis of endocrine disruption in Southern California coastal fish using an aquatic multispecies microarray. *Environ Health Perspect* **117**, 223–230, <https://doi.org/10.1289/ehp.11627> (2009).
26. Baker, M. E. *et al.* Molecular analysis of endocrine disruption in hornhead turbot at wastewater outfalls in southern California using a second generation multi-species microarray. *PLoS One* **8**, e75553, <https://doi.org/10.1371/journal.pone.0075553> (2013).
27. Eden, E., Navon, R., Steinfeld, I., Lipson, D. & Yakhini, Z. GOrrilla: a tool for discovery and visualization of enriched GO terms in ranked gene lists. *BMC Bioinformatics* **10**, 48, <https://doi.org/10.1186/1471-2105-10-48> (2009).
28. Herrero, J. *et al.* Ensembl comparative genomics resources. *Database* **2016**, bav096–bav096, <https://doi.org/10.1093/database/bav096> (2016).
29. Vilella, A. J. *et al.* EnsemblCompara GeneTrees: Complete, duplication-aware phylogenetic trees in vertebrates. *Genome Res* **19**, 327–335, <https://doi.org/10.1101/gr.073585.107> (2009).
30. Davis-Turak, J. *et al.* Genomics pipelines and data integration: challenges and opportunities in the research setting. *Expert Rev Mol Diagn* **17**, 225–237, <https://doi.org/10.1080/14737159.2017.1282822> (2017).
31. Hardiman, G. *et al.* Systems analysis of the prostate transcriptome in African-American men compared with European-American men. *Pharmacogenomics* **17**, 1129–1143, <https://doi.org/10.2217/pgs-2016-0025> (2016).
32. Irish, J. C. *et al.* Amplification of WHSC1L1 regulates expression and estrogen-independent activation of ERalpha in SUM-44 breast cancer cells and is associated with ERalpha over-expression in breast cancer. *Mol Oncol* **10**, 850–865, <https://doi.org/10.1016/j.molonc.2016.02.003> (2016).
33. Xu, E. G. *et al.* Time- and Oil-Dependent Transcriptomic and Physiological Responses to Deepwater Horizon Oil in Mahi-Mahi (*Coryphaena hippurus*) Embryos and Larvae. *Environ Sci Technol* **50**, 7842–7851, <https://doi.org/10.1021/acs.est.6b02205> (2016).
34. Howe, K. *et al.* The zebrafish reference genome sequence and its relationship to the human genome. *Nature* **496**, 498–503, <https://doi.org/10.1038/nature12111> (2013).
35. Schmidt-Heck, W. *et al.* Global transcriptional response of human liver cells to ethanol stress of different strength reveals hormetic behavior. *Alcohol Clin Exp Res*. <https://doi.org/10.1111/acer.13361> (2017).
36. Shepard, J. L. *et al.* A zebrafish bmyb mutation causes genome instability and increased cancer susceptibility. *Proc Natl Acad Sci USA* **102**, 13194–13199, <https://doi.org/10.1073/pnas.0506583102> (2005).
37. Horton, J. D. *et al.* Combined analysis of oligonucleotide microarray data from transgenic and knockout mice identifies direct SREBP target genes. *Proc Natl Acad Sci USA* **100**, 12027–12032, <https://doi.org/10.1073/pnas.1534923100> (2003).
38. Kohjima, M. *et al.* SREBP-1c, regulated by the insulin and AMPK signaling pathways, plays a role in nonalcoholic fatty liver disease. *Int J Mol Med* **21**, 507–511 (2008).
39. Matsuda, M. *et al.* SREBP cleavage-activating protein (SCAP) is required for increased lipid synthesis in liver induced by cholesterol deprivation and insulin elevation. *Genes Dev* **15**, 1206–1216, <https://doi.org/10.1101/gad.891301> (2001).
40. Moon, Y. A. The SCAP/SREBP Pathway: A Mediator of Hepatic Steatosis. *Endocrinol Metab (Seoul)* (2017).
41. Martella, A. *et al.* Bisphenol A Induces Fatty Liver by an Endocannabinoid-Mediated Positive Feedback Loop. *Endocrinology* **157**, 1751–1763, <https://doi.org/10.1210/en.2015-1384> (2016).
42. Renaud, L. *et al.* The Plasticizer Bisphenol A Perturbs the Hepatic Epigenome: A Systems Level Analysis of the miRNome. *Genes (Basel)* **8**, <https://doi.org/10.3390/genes8100269> (2017).
43. Forner-Piquer, I. *et al.* Dose-Specific Effects of Di-Isononyl Phthalate on the Endocannabinoid System and on Liver of Female Zebrafish. *Endocrinology* **158**, 3462–3476, <https://doi.org/10.1210/en.2017-00458> (2017).
44. Hay, N. & Sonenberg, N. Upstream and downstream of mTOR. *Genes Dev* **18**, 1926–1945, <https://doi.org/10.1101/gad.1212704> (2004).
45. Laplante, M. & Sabatini, D. M. mTOR signaling in growth control and disease. *Cell* **149**, 274–293, <https://doi.org/10.1016/j.cell.2012.03.017> (2012).
46. Pusapati, R. V. *et al.* mTORC1-Dependent Metabolic Reprogramming Underlies Escape from Glycolysis Addiction in Cancer Cells. *Cancer Cell* **29**, 548–562, <https://doi.org/10.1016/j.ccell.2016.02.018> (2016).
47. Catania, C., Binder, E. & Cota, D. mTORC1 signaling in energy balance and metabolic disease. *Int J Obes* **35**, 751–761 (2011).
48. Williamson, D. L. *et al.* Altered nutrient response of mTORC1 as a result of changes in REDD1 expression: effect of obesity vs. REDD1 deficiency. *Journal of Applied Physiology* **117**, 246 (2014).

49. Bartolomé, A. *et al.* Pancreatic β -Cell Failure Mediated by mTORC1 Hyperactivity and Autophagic Impairment. *Diabetes* **63**, 2996–3008, <https://doi.org/10.2337/db13-0970> (2014).
50. Blandino-Rosano, M. *et al.* mTORC1 signaling and regulation of pancreatic β -cell mass. *Cell Cycle* **11**, 1892–1902, <https://doi.org/10.4161/cc.20036> (2012).
51. Kenerson, H. L., Subramanian, S., McIntyre, R., Kazami, M. & Yeung, R. S. Livers with constitutive mTORC1 activity resist steatosis independent of feedback suppression of Akt. *PLoS One* **10**, e0117000, <https://doi.org/10.1371/journal.pone.0117000> (2015).
52. Peterson, T. R. *et al.* mTOR complex 1 regulates lipin 1 localization to control the SREBP pathway. *Cell* **146**, 408–420, <https://doi.org/10.1016/j.cell.2011.06.034> (2011).
53. Quinn, W. J. 3rd & Birnbaum, M. J. Distinct mTORC1 pathways for transcription and cleavage of SREBP-1c. *Proc Natl Acad Sci USA* **109**, 15974–15975, <https://doi.org/10.1073/pnas.1214113109> (2012).
54. Laplante, M. & Sabatini, D. M. mTOR signaling at a glance. *J Cell Sci* **122**, 3589–3594, <https://doi.org/10.1242/jcs.051011> (2009).
55. Harding, H. P. *et al.* Ppp1r15 gene knockout reveals an essential role for translation initiation factor 2 alpha (eIF2alpha) dephosphorylation in mammalian development. *Proc Natl Acad Sci USA* **106**, 1832–1837, <https://doi.org/10.1073/pnas.0809632106> (2009).
56. Ramakrishnan, V. M. *et al.* Restoration of Physiologically Responsive Low-Density Lipoprotein Receptor-Mediated Endocytosis in Genetically Deficient Induced Pluripotent Stem Cells. *Sci Rep* **5**, 13231, <https://doi.org/10.1038/srep13231> (2015).
57. Ghodke-Puranik, Y. *et al.* Lupus-Associated Functional Polymorphism in PNP Causes Cell Cycle Abnormalities and Interferon Pathway Activation in Human Immune Cells. *Arthritis & Rheumatology* **69**, 2328–2337, <https://doi.org/10.1002/art.40304> (2017).
58. Iezaki, T. *et al.* The Transcriptional Modulator Interferon-Related Developmental Regulator 1 in Osteoblasts Suppresses Bone Formation and Promotes Bone Resorption. *J Bone Miner Res* **31**, 573–584, <https://doi.org/10.1002/jbmr.2720> (2016).
59. Hamidi, T. *et al.* Nuclear protein 1 promotes pancreatic cancer development and protects cells from stress by inhibiting apoptosis. *J Clin Invest* **122**, 2092–2103, <https://doi.org/10.1172/jci60144> (2012).
60. Castillon, G. A., Watanabe, R., Taylor, M., Schwabe, T. M. & Riezman, H. Concentration of GPI-anchored proteins upon ER exit in yeast. *Traffic* **10**, 186–200, <https://doi.org/10.1111/j.1600-0854.2008.00857.x> (2009).
61. Moon, Y. A., Shah, N. A., Mohapatra, S., Warrington, J. A. & Horton, J. D. Identification of a mammalian long chain fatty acyl elongase regulated by sterol regulatory element-binding proteins. *J Biol Chem* **276**, 45358–45366, <https://doi.org/10.1074/jbc.M108413200> (2001).
62. Lee, A. S. The ER chaperone and signaling regulator GRP78/BiP as a monitor of endoplasmic reticulum stress. *Methods* **35**, 373–381, <https://doi.org/10.1016/j.ymeth.2004.10.010> (2005).
63. Gupta, R. K. *et al.* Di-(2-ethylhexyl) phthalate and mono-(2-ethylhexyl) phthalate inhibit growth and reduce estradiol levels of antral follicles *in vitro*. *Toxicology and applied pharmacology* **242**, 224–230, <https://doi.org/10.1016/j.taap.2009.10.011> (2010).
64. Roh, J. Y., Jung, I. H., Lee, J. Y. & Choi, J. Toxic effects of di(2-ethylhexyl)phthalate on mortality, growth, reproduction and stress-related gene expression in the soil nematode *Caenorhabditis elegans*. *Toxicology* **237**, 126–133, <https://doi.org/10.1016/j.tox.2007.05.008> (2007).
65. Mahankali, M., Peng, H.-J., Cox, D. & Gomez-Cambronero, J. The mechanism of cell membrane ruffling relies on a phospholipase D2 (PLD2), GRB2 and RAC2 association. *Cellular signalling* **23**, 1291–1298, <https://doi.org/10.1016/j.cellsig.2011.03.010> (2011).
66. Klingenspor, M., Xu, P., Cohen, R. D., Welch, C. & Reue, K. Altered gene expression pattern in the fatty liver dystrophy mouse reveals impaired insulin-mediated cytoskeleton dynamics. *J Biol Chem* **274**, 23078–23084 (1999).
67. Langner, C. A. *et al.* The fatty liver dystrophy (fld) mutation. A new mutant mouse with a developmental abnormality in triglyceride metabolism and associated tissue-specific defects in lipoprotein lipase and hepatic lipase activities. *J Biol Chem* **264**, 7994–8003 (1989).
68. Tong, P. *et al.* Insulin-induced cortical actin remodeling promotes GLUT4 insertion at muscle cell membrane ruffles. *J Clin Invest* **108**, 371–381, <https://doi.org/10.1172/jci12348> (2001).
69. Yki-Jarvinen, H. Fat in the liver and insulin resistance. *Ann Med* **37**, 347–356, <https://doi.org/10.1080/07853890510037383> (2005).
70. Yki-Jarvinen, H. Non-alcoholic fatty liver disease as a cause and a consequence of metabolic syndrome. *Lancet Diabetes Endocrinol* **2**, 901–910, [https://doi.org/10.1016/s2213-8587\(14\)70032-4](https://doi.org/10.1016/s2213-8587(14)70032-4) (2014).
71. Lallukka, S. & Yki-Jarvinen, H. Non-alcoholic fatty liver disease and risk of type 2 diabetes. *Best Pract Res Clin Endocrinol Metab* **30**, 385–395, <https://doi.org/10.1016/j.beem.2016.06.006> (2016).
72. Lange, K., Brandt, U., Gartzke, J. & Bergmann, J. Action of insulin on the surface morphology of hepatocytes: role of phosphatidylinositol 3-kinase in insulin-induced shape change of microvilli. *Exp Cell Res* **239**, 139–151, <https://doi.org/10.1006/excr.1997.3894> (1998).
73. Lee, K. Y. *et al.* Bmp2 is critical for the murine uterine decidual response. *Mol Cell Biol* **27**, 5468–5478, <https://doi.org/10.1128/mcb.00342-07> (2007).
74. Lee, K. Y. & DeMayo, F. J. Animal models of implantation. *Reproduction* **128**, 679–695, <https://doi.org/10.1530/rep.1.00340> (2004).
75. Strakovsky, R. S. *et al.* In utero growth restriction and catch-up adipogenesis after developmental di(2-ethylhexyl) phthalate exposure cause glucose intolerance in adult male rats following a high-fat dietary challenge. *J Nutr Biochem* **26**, 1208–1220, <https://doi.org/10.1016/j.jnutbio.2015.05.012> (2015).
76. Carnevali, O. *et al.* DEHP impairs zebrafish reproduction by affecting critical factors in oogenesis. *PLoS One* **5**, e10201, <https://doi.org/10.1371/journal.pone.0010201> (2010).
77. Abdel-Maksoud, F. M., Leasor, K. R., Butzen, K., Braden, T. D. & Akingbemi, B. T. Prenatal Exposures of Male Rats to the Environmental Chemicals Bisphenol A and Di(2-Ethylhexyl) Phthalate Impact the Sexual Differentiation Process. *Endocrinology* **156**, 4672–4683, <https://doi.org/10.1210/en.2015-1077> (2015).
78. Wu, S. *et al.* Dynamic effect of di-2-(ethylhexyl) phthalate on testicular toxicity: epigenetic changes and their impact on gene expression. *Int J Toxicol* **29**, 193–200, <https://doi.org/10.1177/1091581809355488> (2010).
79. Isenberg, J. S. *et al.* Reversibility and persistence of di-2-ethylhexyl phthalate (DEHP)- and phenobarbital-induced hepatocellular changes in rodents. *Toxicol Sci* **64**, 192–199 (2001).
80. Martinez-Arguelles, D. B., Culty, M., Zirkin, B. R. & Papadopoulos, V. In utero exposure to di-(2-ethylhexyl) phthalate decreases mineralocorticoid receptor expression in the adult testis. *Endocrinology* **150**, 5575–5585, <https://doi.org/10.1210/en.2009-0847> (2009).
81. Ge, R. S. *et al.* Stimulation of testosterone production in rat Leydig cells by aldosterone is mineralocorticoid receptor mediated. *Mol Cell Endocrinol* **243**, 35–42, <https://doi.org/10.1016/j.mce.2005.08.004> (2005).
82. Brun, H. P. *et al.* Pig Leydig cell culture: a useful *in vitro* test for evaluating the testicular toxicity of compounds. *Toxicology and applied pharmacology* **108**, 307–320 (1991).
83. Taghizadeh, M. & Noruzinia, M. Lovastatin Reduces Stemness via Epigenetic Reprogramming of BMP2 and GATA2 in Human Endometrium and Endometriosis. *Cell Journal (Yakhteh)* **19**, 50–64 (2017).
84. Fu, B. *et al.* Epigenetic regulation of BMP2 by 1,25-dihydroxyvitamin D3 through DNA methylation and histone modification. *PLoS One* **8**, e61423, <https://doi.org/10.1371/journal.pone.0061423> (2013).
85. Corradetti, B. *et al.* Bis-(2-ethylhexyl) phthalate impairs spermatogenesis in zebrafish (*Danio rerio*). *Reprod Biol* **13**, 195–202, <https://doi.org/10.1016/j.repbio.2013.07.003> (2013).
86. Golshan, M. *et al.* Di-(2-ethylhexyl)-phthalate disrupts pituitary and testicular hormonal functions to reduce sperm quality in mature goldfish. *Aquat Toxicol* **163**, 16–26, <https://doi.org/10.1016/j.aquatox.2015.03.017> (2015).

87. Sasik, R., Woelk, C. H. & Corbeil, J. Microarray truths and consequences. *J Mol Endocrinol* **33**, 1–9 (2004).
88. Cole, S. W., Galic, Z. & Zack, J. A. Controlling false-negative errors in microarray differential expression analysis: a PRIM approach. *Bioinformatics* **19**, 1808–1816 (2003).
89. Andrews, S. FastQC: a quality control tool for high throughput sequence data. (2010).
90. Martin, M. Cutadapt removes adapter sequences from high-throughput sequencing reads. **17**, <https://doi.org/10.14806/ej.17.1.20010-12> (2011).
91. Kim, D. *et al.* TopHat2: accurate alignment of transcriptomes in the presence of insertions, deletions and gene fusions. *Genome Biol* **14**, R36, <https://doi.org/10.1186/gb-2013-14-4-r36> (2013).
92. Anders, S., Pyl, P. T. & Huber, W. HTSeq—a Python framework to work with high-throughput sequencing data. *Bioinformatics* **31**, 166–169, <https://doi.org/10.1093/bioinformatics/btu638> (2015).
93. Love, M. I., Huber, W. & Anders, S. Moderated estimation of fold change and dispersion for RNA-seq data with DESeq. *2. Genome Biol* **15**, 550, <https://doi.org/10.1186/s13059-014-0550-8> (2014).
94. Trapnell, C. *et al.* Differential gene and transcript expression analysis of RNA-seq experiments with TopHat and Cufflinks. *Nat Protoc* **7**, 562–578, <https://doi.org/10.1038/nprot.2012.016> (2012).
95. Eden, E., Lipson, D., Yogev, S. & Yakhini, Z. Discovering motifs in ranked lists of DNA sequences. *PLoS Comput Biol* **3**, e39, <https://doi.org/10.1371/journal.pcbi.0030039> (2007).
96. Supek, F., Bošnjak, M., Škunca, N. & Šmuc, T. REVIGO Summarizes and Visualizes Long Lists of Gene Ontology Terms. *PLOS ONE* **6**, e21800, <https://doi.org/10.1371/journal.pone.0021800> (2011).
97. Chen, J., Xu, H., Aronow, B. J. & Jegga, A. G. Improved human disease candidate gene prioritization using mouse phenotype. *BMC Bioinformatics* **8**, 392, <https://doi.org/10.1186/1471-2105-8-392> (2007).
98. Chen, J., Aronow, B. J. & Jegga, A. G. Disease candidate gene identification and prioritization using protein interaction networks. *BMC Bioinformatics* **10**, 73, <https://doi.org/10.1186/1471-2105-10-73> (2009).
99. Chen, J., Bardes, E. E., Aronow, B. J. & Jegga, A. G. ToppGene Suite for gene list enrichment analysis and candidate gene prioritization. *Nucleic Acids Res* **37**, W305–311, <https://doi.org/10.1093/nar/gkp427> (2009).

Acknowledgements

We thank Drs. Bethany J. Wolf, Jeremy L. Barth and Russell A. Norris for useful discussions. We thank Dr. Roman Sasik and Dr. E.S. Hazard for discussions on the bioinformatics analyses. GH acknowledges funding from SC EPSCoR and start-up funding from College of Medicine at the Medical University of South Carolina.

Author Contributions

M.H. and G.H. performed the experiments reported here, contributed to the scientific direction of this study, and prepared the manuscript, all under the guidance of G.H., who originally conceived the study. M.H., W.D.S., L.R., O.C. and G.H. performed data analyses. All authors, reviewed and edited the manuscript.

Additional Information

Supplementary information accompanies this paper at <https://doi.org/10.1038/s41598-018-20266-8>.

Competing Interests: The authors declare that they have no competing interests.

Publisher's note: Springer Nature remains neutral with regard to jurisdictional claims in published maps and institutional affiliations.



Open Access This article is licensed under a Creative Commons Attribution 4.0 International License, which permits use, sharing, adaptation, distribution and reproduction in any medium or format, as long as you give appropriate credit to the original author(s) and the source, provide a link to the Creative Commons license, and indicate if changes were made. The images or other third party material in this article are included in the article's Creative Commons license, unless indicated otherwise in a credit line to the material. If material is not included in the article's Creative Commons license and your intended use is not permitted by statutory regulation or exceeds the permitted use, you will need to obtain permission directly from the copyright holder. To view a copy of this license, visit <http://creativecommons.org/licenses/by/4.0/>.

© The Author(s) 2018