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Oribatid mites show how climate and latitudinal gradients in organic matter can drive large-scale biodiversity patterns of soil communities

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3 **Oribatid mites show how climate and latitudinal gradients in organic matter can drive**
4 **large-scale biodiversity patterns of soil communities**

5

6 **Short running title:** *Macroecological determinants of soil animals*

7

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28 coordinating and conducting the 1998 Countryside Survey and subsequent processing of
29 samples. We thank three anonymous reviewers for their constructive comments.

30

Abstract

Aim: The factors determining spatial distributions and diversity of terrestrial invertebrates are typically investigated at small scales. Large-scale studies are ~~particularly missing~~ lacking for soil animals, which control microbial communities and represent one of the most diverse yet poorly known animal assemblages. Here, we analyzed a major group (Oribatida) to test whether belowground macroecological patterns can be predicted by climatic variables, vegetation, and large-scale variation in key soil properties.

Location: ~~we~~ We modelled the multivariate distribution of more than 100 species using biodiversity data collected across Great Britain in the framework of the Countryside Survey (<http://www.countrysidesurvey.org.uk>).

Methods: We analysed species-level data from 582 samples collected across 162 hectads (10 × 10 km) covering the largest possible range of vegetation types, soil properties and climatic conditions within GB. We created the first large-scale maps of soil animal diversity metrics at the GB scale, including novel estimates of metrics of phylogenetic diversity. Using structural equation modelling, we quantified the direct and indirect effects of location (latitude, longitude), plant community structure, and abiotic factors such as precipitation on species composition, richness, and phylogenetic diversity.

Results: We found that variation in species composition follows a latitudinal gradient with diversity generally increasing northward. The latitudinal variation in species composition drives phylogenetic diversity via changes in both species richness and phylogenetic distance between species. This gradient is mostly determined by latitudinal variation in precipitation and organic matter, which were very good predictor of species composition. Precipitation and organic matter were, however, relatively weak while statistically significant predictors of diversity metrics.

Conclusions: Past studies have emphasized the unpredictability of species distributions and variation in species composition in hyper diverse soil animal communities. However, past studies were conducted at small scales, where stochastic factors may weaken the signal of deterministic factors. Oribatid mites in our study show for the first time that large scale latitudinal gradients in climate and organic matter predict not only variation in species composition but also taxonomic and phylogenetic diversity of soil animal communities.

Keywords: soil macroecology; animals; diversity; distribution; community phylogenetics; Oribatida

1. Introduction

In the last two decades, one of the major goals of ecology has been to understand the relative roles of the many factors that structure ecological communities in space and time but the majority of studies have focused on aboveground communities, particularly plant communities (Chesson, 2000; Clark & McLachlan, 2003; Hubbell, 2005; Levine & HilleRisLambers, 2009). More recently, ecologists have started to better investigate the interactions between aboveground and belowground communities and how these interactions drive the composition and diversity of both communities (Kardol et al., 2006; Van Der Heijden et al., 2008; de Vries et al., 2012; Prober et al., 2015). Traditionally, most studies investigating aboveground-belowground linkages have been conducted at relatively local spatial scales but some regional and global scale analyses of soil microbial communities have shed light on the large scale determinants of these communities (Fierer et al., 2009; de Vries et al., 2012; Ramirez et al., 2014). Local and fine scale variation seems mostly due to interactions determined by the patchy distribution of resources and plant species (Bezemer et al., 2010; Thomson et al., 2010) while spatial gradients in vegetation types and abiotic factors such as pH and climatic conditions are the major correlates of microbial distribution at regional scales (Fierer et al., 2009; Griffiths et al., 2011). Protists, too, follow similar macroecological patterns (Soininen, 2011). For some groups such as arbuscular mycorrhizal fungi (AMF), global studies have started to reveal the relative roles of local, regional and historical factors on community structure and diversity (Davison et al., 2015) but for soil animals large-scale studies are missing despite some synthesis data having provided insight in the macroecology of soil arthropods, nematodes and oligochaetes (Decaëns, 2010; Brusaard et al., 2012; Nielsen, 2014). Studies focusing on selected assemblages at relatively local scales (Lindo & Winchester, 2009; Nielsen et al., 2010; Caruso et al., 2011) have shown that, similarly to microbial communities, soil animal communities are structured at multiple spatial scales, with many species being dispersal-limited over certain scales (Ettema & Wardle, 2002; Wardle, 2006) and soil environmental

90 heterogeneity being high already at very small scale (e.g. <100cm; Ettema & Wardle 2002).
91 This small scale heterogeneity promotes community diversity in both animals and microbes
92 and is mostly due to small scale variation in soil properties such as pH, the concentration of
93 organic matter and key nutrient such as P and N, and also structural variation in soil such as
94 variation in the physical distribution of aggregates (Dumbrell et al., 2010; Nielsen et al.,
95 2010). Nevertheless, much spatial variation in the structure and diversity of soil communities,
96 especially animals, is often left unexplained by variation in environmental variables or other
97 biotic factors. Previous studies have hypothesised that stochastic population dynamics,
98 including dispersal limitation, may sometimes play a major role in this variation (Lindo &
99 Winchester, 2009; Caruso et al., 2011). At large scales, the few studies available on soil
100 animals have concentrated on classical macroecological patterns such as species-area
101 relationships, altitudinal and latitudinal gradients in diversity and some insight on patterns of
102 phylogenetic diversity (see review in Brusaard et al., 2012) but most datasets focused on the
103 highest taxonomic ranks (e.g., family or class level; Nielsen, 2014) or, as noted by Decaëns,
104 (2010) are biased in terms of sampling efforts towards temperate countries, and in general
105 lack the resolution necessary to disentangle the contribution of multiple factors at multiple
106 spatial scales. The only quantitative study on the regional variation of soil animal
107 communities at the species level (Zaitsev *et al.* 2013) showed that studies conducted at
108 relatively small scales cannot capture the long-term effects of the historical processes that
109 contribute to large scale gradients in species richness and community composition. Overall,
110 large-scale studies that help disentangle the roles of abiotic and biotic factors that structure
111 soil communities at regional scales are in their infancy.

112 Here we focused on a unique dataset of the species of oribatid mites collected during the
113 first assessment of soil biodiversity across Great Britain undertaken in 1998 (known as the
114 GB Countryside Survey or CS: <http://www.countrysidesurvey.org.uk>). This survey produced
115 a baseline dataset across all major soil types and habitats (Black et al., 2002) and showed
116 that populations of microbes and microarthropods varied across major environmental zones,

117 vegetation classes and soil types (Black et al., 2003; Griffiths et al., 2011; Keith et al., 2015).
118 Oribatid mites (Acari, Acariformes) are a cornerstone of soil food webs worldwide: over
119 10,000 species have been described and they can reach densities of up to 400,000 ind./m²
120 in forest soils, although they occur in all biomes including continental Antarctica (Coleman et
121 al., 2004). Oribatid mites are one of the most ancient groups of terrestrial animals and have
122 been part of the soil food webs ever since soil have appeared on the geological record about
123 400 mya (Shear et al., 1984; Norton et al., 1988). They appear for the first time in the fossil
124 record of the Devonian site of Rhynie Chert (407-397 mya, Aberdeenshire, Scotland)
125 although a relatively recent molecular clock suggests a much earlier origin (Schaefer et al.,
126 2010). For all these reasons, oribatid mites provide an excellent model to analyze the role of
127 abiotic and biotic factors in structuring diversity and composition of belowground animals at
128 regional scales. We used the dataset of oribatid mites to conduct a species-level analysis of
129 the determinants of community structure and diversity of this major group of soil animals.
130 The CS is a unique audit of vegetation, soils, habitats and landscape across GB that began
131 in 1978 (Firbank et al., 2003; Keith *et al.* 2015). Using a spatially explicit approach, we
132 created the first maps of diversity metrics at the GB scale for a major soil animal group,
133 including the first estimates of phylogenetic diversity (Faith, 1992; Cadotte et al., 2010) and
134 used structural equation modelling (Grace, 2006) to estimate the direct and indirect effects of
135 location (latitude, longitude), abiotic factors such as precipitation, and plant community
136 structure on oribatid mite species composition and diversity. We hypothesized that large-
137 scale gradients in this belowground community are directly driven by large-scale gradients in
138 abiotic factors (e.g., climatic variables) but also via the effects of these factors on plant
139 community structure and edaphic properties such as organic matter.

140

141

142

143 **Methods**

144 ~~Data~~~~base~~ *Background background*

145 The data analysed in this study were collected in the framework of the Country Survey audit
146 (Firbank et al., 2003; Keith *et al.* 2015). This environmental audit is based on a stratification
147 of GB into land classes, each land class being characterised by a combination of climate,
148 altitude and location (Firbank et al., 2003; Sheail & Bunce, 2003). Each sampling location
149 was assigned ~~a Broad Habitat (BH) and~~ an Aggregate Vegetation Class (AVC). ~~BH is a~~
150 ~~classification consisting of 27 habitats that are used in the Land Cover Map accounting for~~
151 ~~the entire land surface of GB, and~~; AVC is a high-level grouping of vegetation types
152 produced from a classification of plant communities from the original CS vegetation plots and
153 includes eight categories ~~that group common habitats~~ (crops and weeds, tall grass and herb,
154 fertile grassland, infertile grassland, lowland woodland, upland woodland, moorland-grass
155 mosaic, and, heath and bog (Bunce et al., 1999). Specifically, we analysed 582 samples
156 mostly collected between 29 May and 28 October in 1998 with some samples collected
157 between June and August in 1999. The samples analysed in this study were collected
158 across 162 10 x 10 km plots (hectads), with an average of 4 locations sampled within each
159 hectad. Each sample location was associated with information on vegetation, soil properties
160 and land-use, produced during CS. For the collection of each soil sample, surface vegetation
161 was removed leaving the litter layer intact and a soil core (4 cm diameter, 8 cm depth) was
162 taken. Cores were placed immediately in cool boxes and sent to the laboratory at the Centre
163 for Ecology & Hydrology Lancaster for extraction of invertebrates.

164

165 2.2 *Oribatid extraction and identification*

166 Cores were processed over five days using a dry Tullgren extraction method and all
167 invertebrate specimens collected into 70% ethanol preservative (Emmett et al., 2010) . Once
168 collected, the soil invertebrates were identified to broad taxa, separated and counted under a

169 stereomicroscope. Specimens of Acari (mites) from each extract were removed into another
170 vial and sent for separation and identification of oribatid mites species. Specimens were
171 identified at $\times 400$ magnification and, where necessary, were cleared for 24h using lactic acid
172 at room temperature before being mounted in glass cavity slides. The unpublished
173 monograph of British oribatid mites by M. Luxton and other specialist primary literature were
174 used to identify oribatid mites to the species level; identifications have since been checked
175 against Weigmann, (2006). Weigmann (2006) plus several other specialist publications were
176 also used to define geographic distributions and ecological traits of the taxa. Plot codes,
177 ~~Oribatid-oribatid~~ species records and taxonomic details were then collated into a dataset ~~that~~
178 ~~is available upon request from the NERC Environmental Information Centre (see Keith et al.~~
179 ~~2018 for details to access the data).~~

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181 2.3 Associated environmental data

182 Existing ~~data on soil (Black et al. 2016), and vegetation (Barr et al. 2014) and habitat data~~
183 ~~from CS were collated for the 582 plots with oribatid mite records samples of soil fauna and~~
184 ~~are available upon request from the NERC Environmental Information Centre (see Barr et al.~~
185 ~~2014 and Black et al. 2016 for details to access the vegetation and habitat data).~~ Soil
186 properties were collected from a separate core taken adjacent to the core used to extract soil
187 animals; soil data included moisture content, pH, organic matter (loss-on-ignition), total C
188 content and total N content. The sampling protocol and detailed methods used for these soil
189 analyses can be found in Emmett et al. (2008) and the data are reported in more detail
190 elsewhere (Reynolds et al., 2013). For vegetation composition, ordination scores were used
191 from the first three axes of a Detrended Correspondence Analysis (DCA) using binary plant
192 species data from the same plots.

193 Climate data associated with each sampling plot was derived from the UKCP09: Met Office
194 gridded land surface climate observations at 5×5 km resolution (Met Office, 2017). These

195 data were used to calculate average values of mean annual temperature and mean annual
196 rainfall for the period 1992–1997, in order to incorporate recent climatic trends.

197

198 2.4 Statistical analyses

199 Community and environmental data

200 Records of oribatid mites across sampling locations were collated at 10km × 10km/hectad
201 resolution for a total of 162 10 x 10 km squares, and the associated environmental data were
202 averaged at this resolution (~~see Barr et al. 2014, Black et al. 2016 and Keith et al. 2018 to~~
203 ~~access the data from NERC Environmental Information Centre).~~

204 We used the spatial interpolation method of kriging (Matheron, 1963; Wagner, 2003) to
205 illustrate spatial variation in community structure and metric of diversity (see below for the
206 metrics used). The spatial structure of the variables was quantified with the empirical
207 semivariogram (Wagner, 2003; Bivand et al., 2008) and then fitted with a theoretical
208 variogram model (i.e., exponential or Gaussian, or spherical models) to estimate values at
209 unmeasured locations. We used the R library "~~geoR~~", "~~maps~~", "~~mapdata~~" and "~~gstat~~"
210 for calculating variograms, kriging estimation and mapping of results. (~~See also~~ Bivand et al.
211 ~~(2008)~~ for further details.

212 We used a multivariate regression approach based on Principal Coordinate Analysis
213 (Legendre & Legendre, 1998; Borcard et al., 2004) to quantify the relative importance of
214 location (space) and environment (temperature, plant community composition, pH,
215 precipitation, organic matter) on oribatid mite community structure. PCoA was applied to the
216 Jaccard distance matrix obtained by the presence-absence distribution of species, and a
217 distance based RDA (db-RDA) was used to estimate the effect of space and environment on
218 the multivariate distribution of species.

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219 To quantify the effect of “space” (i.e., location), we used latitude, longitude and the method
220 of principal coordinate analysis of neighbour matrices (PCNM; Borcard & Legendre, 2002),
221 which defines a set of spatial factors that parsimoniously account for patterns in species
222 distribution at multiple scales. The final set of PCNM vectors was defined using a
223 multivariate extension of the Akaike information criterion (AIC; Dray et al., 2006). Variance
224 partitioning was computed to estimate the amount of fraction uniquely attributable to space
225 and environment, and the variation shared between space and environment (Borcard et al.,
226 1992; Legendre & Legendre, 1998). Besides observed species number per hectad we also
227 calculated species rarefaction curve (Gotelli & Colwell, 2001) for each hectad and estimated
228 the hectad asymptotic richness using the Chao estimator (O’Hara 2005; Chiu *et al.* 2014). All
229 multivariate analyses and estimates of species richness were performed using the R
230 package ‘vegan’ “~~vegan~~” (Oksanen et al., 2007).

231

232 Phylogenetic methods

233 The phylogenetic tree was reconstructed based on 18S rDNA. Sequences were downloaded
234 from NCBI (<https://www.ncbi.nlm.nih.gov>) or, if not available, were newly generated
235 sequenced at the J.F. Blumenbach Institute of Zoology and Anthropology, University of
236 Göttingen.

237 Genomic DNA was extracted from single individuals using the DNeasy® Blood and Tissue
238 Kit (Qiagen, Manchester, UK) following the manufacturer’s protocol for animal tissue.
239 Amplification of the 18S region was performed in 25 µl volumes containing 12.5 µl
240 HotStarTaq Mastermix (Qiagen), 5 µl of template DNA, 1 µl of each primer (100 pM) and 5.5
241 µl H₂O. Primers for PCR were 5’ -TAC- CTGGTTGATCCTGCCAG-3’ (forward) and 5’ -
242 TAATGATCCTTCCGC AGGTTAC-3’ (reverse) (Domes et al., 2007). The PCR protocol
243 consisted of an initial activation step at 95 °C for 15 min, 35 amplification cycles (95°C for 45
244 s, 57° C for 60 s, 72°C for 60 s) and a final elongation step at 72 °C for 10 min. All PCR

245 products were visualized on a 1% agarose gel, purified with the QIAquick PCR Purification
 246 Kit (Qiagen), and sequenced by Microsynth Seqlab (Göttingen, Germany), using the
 247 additional sequencing primers 18S554f 5'-AAGTCTGG TGCCAGCAGCCGC-3' , 18S1282r
 248 5'-TCACTCCACCAACTA AGAACGG C-3' , 18S1150f 5' -
 249 ATTGACGGAAGGGCACCACCAG-3' and 18S614r 5'- TCCAACACGAGCTTTTAAACC-3'
 250 (Domes et al., 2007). In total, we used 51 species for the phylogenetic tree, including four
 251 outgroup taxa. All taxa and accession numbers are available at GenBank (Supporting
 252 Information, Appendix S1, Table S1). We aimed to represent each family in the GB dataset
 253 with at least one species but very few rarer species could not be represented either because
 254 sequences are not available in public database or because we did not have sufficient
 255 material to sequence them. In total, the dataset represents 31 out of the 34 families found in
 256 the GB dataset and the three families we could not represent were very rare and present
 257 with very low abundances. Sequences were assembled and ambiguous positions were
 258 corrected in Sequencher 5.3 (Gene Codes Corporation, Ann Arbor, Michigan, USA) and
 259 aligned using ClustalW implemented in BioEdit v7.0.1 (Hall, 1999) with the multiple
 260 alignment parameters gap opening = 30 and gap extension = 0.3.

261 The alignment was truncated to the shortest sequence resulting in a length of 1,743 bp
 262 including gaps. Evolutionary model parameters were determined with jModelTest v2.0
 263 (Guindon & Gascuel, 2003; Darriba et al., 2012) using the AIC. The best-fit model for
 264 sequence evolution for 18S was GTR + I + G. The phylogenetic tree was constructed in R
 265 using packages '[ape](#)' (Paradis et al., 2004) and '[phangorn](#)' "~~ape~~" and "~~phangorn~~" (Paradis et
 266 ~~al., 2004~~; (Schliep 2011, Schliep et al., 2017) using Maximum Likelihood and 1,000
 267 bootstrap replicates. The phylogenetic tree was reduced to 31 oribatid mite taxa
 268 representing one species per family using the drop.tip function, the R script is provided in the
 269 Appendix S1 (Supporting Information)

270 We used this oribatid mite phylogenetic tree to calculate metrics of phylogenetic diversity.
 271 The resolution of the available phylogenetic information constrained us to calculate these

metrics at the family level. Specifically, we calculated the Faith's Index PD (Faith, 1992) and two distance based metrics: the mean phylogenetic distance (MPD) and the mean nearest taxon distance (MNTD). The index PD estimates the phylogenetic diversity of a community as the sum of the tree branch lengths connecting all species in the assemblage and as such it can be considered an estimate of point diversity with two components: species richness and amount of phylogenetic information across all the species in the assemblage. The indices MPD and MNTD measure the average phylogenetic distance between species in an assemblage. The MPD is based on mean distance of any taxon from every other taxon while MNTD is average of the distance between any taxon and its closest relative. Community phylogenetic metrics were calculated using the R package "picante" and other packages ('ape' and 'phytools') that support phylogenetically informed statistical analyses (~~'ape', 'phylobase', 'adephylo', 'phytools'~~ ~~"ape", "phylobase", "adephylo", "phytools"~~; Kembel Paradis et al., 2004, Kembel 2010, Revell 2012, Swenson, 2014)

Structural equation modelling (SEM)

In order to quantify direct and indirect effects of climatic and soil variables (temperature, plant communities, pH, precipitation, organic matter) on community species composition (PCoA ordination axes) and diversity (richness and metric of phylogenetic diversity) we used Structural Equation Modelling (Grace, 2006). We started from an *a priori* model (Appendix S2, Fig. S1 in the Supporting Information) assuming that latitude, longitude and elevation drive the spatial variation of climate (i.e. precipitation and temperature), which correlates with spatial variation in organic matter. The variation in climate and organic matter then drives spatial variation in oribatid mite species composition. However, other factors that may vary with latitude and longitude, including biogeographical factors, may drive the spatial distribution of oribatid species. Biogeographical factors, which are implicitly accounted for by latitude and longitude, include the major spatially structured features of the geology of Britain

298 (Toghil, 2005), with northern areas (e.g, Scottish highlands) being generally older but also
299 more affected by last glacial maxima than southern areas (e.g., southeast England).
300 Eventually, all these environmental and geological changes in space determined spatial
301 variation both in species composition and metrics of diversity, including species richness and
302 phylogenetic diversity metrics that combined both richness and compositional information.
303 Starting from this conceptual model, we fitted various versions of the model to the data until
304 we obtained a parsimonious model that could adequately fit the data. Model fit was
305 evaluated using the Chi-square test, and the RMSEA and CFI index, while amount of
306 explained variation in community metrics and diversity indices (R-square) was used to
307 measure the predictive power of the model (Grace, 2006; Shipley, 2016). SEM was
308 performed using the R package 'lavaan' ~~lavaan~~ (Rosseel, 2011).

310 Results

311 A total of 141 species were found in this study, which represented more than one third of
312 known oribatid mites in the British Isles (Luxton, 1996) and the vast majority of belowground
313 species (the CS survey specifically focused on soil species while oribatid mites also live on
314 aboveground moss and tree canopy). Observed species richness ranged from 1 to 28 while
315 the Chao's estimator ranged from 1 to 161 species. Hectads with very low species richness
316 always included arable and very infertile grassland soil, where environmental conditions
317 typically supported only very poor oribatid mite communities or no oribatids at all. These
318 soils were colonised only by very few opportunistic species such as some of the species in
319 the genera *Tectocepheus*, *Liochthonius* and *Pantelozetes*. On the contrary, hectads with
320 high species richness tended to be characterised by woodland or organic soils, where
321 oribatid mites are known to be abundant. Observed hectad species richness displayed both
322 clear latitudinal and longitudinal gradients with hectads in north-west Scotland being richer
323 than in south-east England (Fig. 1). Instead, Chao's estimator showed a very patchy

324 distribution suggesting the existence of hotspots of species richness, mostly located in
 325 central and northern England, and Scotland (Fig. 1).

326 The species-area relationship was best fitted by a sigmoidal model (Fig. 2) meaning the
 327 existence of an upper limit below which species richness is relatively, but not completely,
 328 independent of area. Also, as area increases species richness is predicted to reach an upper
 329 asymptotic level (Lomolino, 2000). The classical power model and semi-log model provided
 330 a much poorer fit to the data.

331 The variation in oribatid mite species composition was mostly driven by the covariation
 332 between organic matter (measured as Loss-on-ignition), pH, and precipitation and
 333 variation in plant species composition, which was almost collinear with amount of organic
 334 matter. Although both community structure and environmental variables follow clear
 335 latitudinal patterns (Fig. 3), the total amount of variance accounted for by measured
 336 environmental variables was only 8%. Yet, this fraction of variation was statistically
 337 significant at $P < 0.05$. There was also 6% of variance accounted for by the spatially
 338 structured effect of environmental and plant variables. The pure effect of latitude, longitude
 339 and PCNMs (i.e., after removing environment) accounted only for 1% of community
 340 variance, meaning that the observed spatial variation in the assemblage is mostly co-varying
 341 with the spatial structure observed in the environmental variables (6%).

342 Metrics of phylogenetic diversity showed different types of spatial patterns (Fig. 4). The
 343 Faith's index showed gradients that were highly correlated to the same ones observed for
 344 plot species richness (compare Fig. 4 with Fig. 1) while MPD and MNTD mostly reflected
 345 longitudinal gradients. MPD is higher in the North and the East while MNTD seems more
 346 variable and reaching the highest value in the South-East (Fig. 4).

347 Structural equation modelling indicated that models including just latitude as a descriptor of
 348 position generally outperformed models with both latitude and longitude in terms of global fit
 349 metrics. For example, all models with both latitude and longitude resulted in Chi-square with

p-values much lower than 0.05 (i.e., model rejected) and with very poor CFI (<0.9) and RMSEA (>0.2) values. We therefore retained latitude and removed longitude from the subsequent models. Although latitude could affect indices of diversity both directly and indirectly, models with a direct link between latitude and diversity indices returned very poor global fit metrics and were therefore not considered further. Details on the models considered during the SEM exercise and their performances are provided Appendix S2. The optimal model (Fig. 5) suggests that organic matter is the major driver of oribatid mite community composition and that variation in species composition determines metrics of phylogenetic diversity. Specifically, greater shifts in oribatid mite community assemblages towards that typical of heath, bog and highly organic soil, were associated with higher phylogenetic diversity (positive correlation between PCoA1 and PD) but also lower mean nearest taxon distance (negative correlation between PCoA1 and MNTD). However, there is also an indirect positive effect of PCoA1 on MNTD via PD (positive correlation between PD and MNTD). The model could account for 50, 16 and 5 % of variance in PCoA 1 (major changes in species composition), PD and MNTD, respectively. The full ~~lavaan~~ output of the SEM is in Appendix S2 (Supporting Information)

366

367 4. Discussion

Soil animal assemblages tend to be very species rich even at small scales. This has been explained as an effect of the high environmental and microbiological heterogeneity that some soil can display already from the 10 m to the sub-metre scale (Anderson, 1975; Giller, 1996; Ettema & Wardle, 2002; Nielsen et al., 2010). A surprisingly large fraction of the variation observed in the distribution of soil species is very often left unexplained by variation in key soil variables such as pH and organic C, or even pollutants (Maraun & Scheu, 2000; Caruso et al., 2011, 2017; Maaß et al., 2015). Also, high degrees of stochasticity seem to characterise assembly dynamics of soil animals such as oribatid mites and collembolans at least at small to medium scales (Maaß et al., 2014; Dirilgen et al., 2018). Still, species

distributions seem structured at small and medium scales even when spatial structure cannot be explained by spatial gradients in environmental variables (Caruso et al., 2011; Zaitsev et al., 2013). At the regional scale of the Netherlands, Zaitsev *et al.* (2013) found that oribatid mite communities significantly changed along the East-West direction in the absence of a significant variation in precipitation and mean annual temperature. However, geological age (bedrock) and amount and quality of organic matter did change from East to West supporting richer communities in the older forest sites (Zaitsev *et al.* 2013). Our dataset supports this idea at the much broader scale of Great Britain, which is characterised by a relationship between climatic gradients and organic matter: in Britain very organic rich soils (i.e. bogs and peatlands) are ~~mostly found~~predominant in the North and West, and are characterised by a colder winter climate with more precipitation. Thus, as mean annual precipitation increases with latitude so does organic matter. This is reflected in our data by statistically significant, latitudinal changes in the oribatid mite communities, which prefer organic soil and woodland over low fertile grassland and cropland. Land use could also contribute to these patterns because, in GB, land is generally much more exploited for intensive farming in the south (e.g., England) than the north (Highlands in Scotland). However, our analysis independently accounted for vegetation types and latitudinal gradients in other properties and our results suggest a prominent role of organic matter *per se*. That means that, given the same land use and vegetation type, sites with higher organic matter are associated to specific oribatid mite composition and higher diversity overall. Species richness and metrics of phylogenetic diversities, too, follow this latitudinal gradient in community structure although metrics of phylogenetic diversity that take into account phylogenetic distance between species (MPD and MNTD) show patterns more complex than just a latitudinal gradient. The SEM showed that variation in distance based metrics of phylogenetic diversity (e.g., MNTD) seemed mostly explained by latitudinal changes in species composition rather than accumulation of species richness and phylogenetic diversity (PD). In fact, the direct and negative effect of the latitudinal changes in species composition on MNTD was statistically significant while the direct and positive effect of PD on MNTD was

405 not. The negative correlation between the latitudinal gradient in oribatid mite composition
 406 and MNTD suggests that the more the oribatid community moves to the species composition
 407 typical of woodland and highly organic soils the less the phylogenetic distance is between a
 408 species and its closest relatives in the local assemblages. This result suggests a process of
 409 environmental filtering and convergence toward specific assemblages (Webb, 2000). The
 410 SEM, however, could explain only 5% of the variance observed in MNTD and 16% of the
 411 variance observed in PD suggesting that the measured environmental variables are
 412 generally weak predictors of these broad biodiversity metrics. On the contrary, the SEM
 413 could explain about 50% of the variance observed in the latitudinal gradient in species
 414 composition, which implies species composition is much more predictable than compound
 415 metrics of biodiversity such as phylogenetic diversity (PD). Specifically, the latitudinal
 416 changes in species composition seem best explained by latitudinal variation in organic
 417 matter and precipitation, regardless of variation in phylogenetic diversity.

418 Latitude directly correlates with precipitation and organic matter distribution merely because
 419 of the north-south climatic gradient. When taking into account the direct and indirect effects
 420 of latitude, precipitation and organic matter on oribatid mite species composition, the
 421 strongest effect was that of organic matter. Precipitation, too, had a statistically significant,
 422 direct effect on community structure but the effect was much smaller than that of organic
 423 matter, which is consistent with Zaitsev *et al.* (2013). Instead, the direct effect of latitude on
 424 species composition was small and not statistically significant, which implies that latitudinal
 425 changes in species composition are driven by latitudinal changes in other variables, namely
 426 precipitation and organic matter. Alternative SEMs that linked latitude, longitude, organic
 427 matter and precipitation directly to metrics of diversity had a very poor global fit supporting
 428 the notion that large-scale gradients in soil oribatid mite diversity are driven by the factors
 429 that drive changes in species composition. Still, changes in species composition explained
 430 only a relatively small fraction of changes in species richness and phylogenetic diversity,
 431 suggesting a potential role for smaller scale heterogeneity. This heterogeneity is not

432 captured by our predictors and suggests that microscale variation in edaphic properties
433 remain a fundamental driver of species distribution and diversity in these communities. This
434 is confirmed by the fact that some hectads resulted to be biodiversity hotspot in terms of
435 estimated species richness. We could not resolve the variables driving this patchy pattern
436 but we speculate that this is driven by soil environmental heterogeneity within hectads, which
437 could be caused by unmeasured variation in habitat fragmentation and land-use intensity
438 (see also supplementary results in the Supporting Information, Appendix S3, Table S2)

439 Despite the latitudinal patterns observed in oribatid mites and contrary to what has been
440 observed in small- and medium-scale studies (Caruso et al., 2011; Maaß et al., 2015), the
441 investigated community had limited spatial structure, even when considering spatial variation
442 that is not explainable by spatial structure in environmental variables. In comparison, the
443 microbial communities of GB seem to be much more spatially structured (Griffiths et al.,
444 2011), which suggests the interesting hypothesis of a decoupling between large-scale
445 patterns in soil microbes and animals.

446

447 **Conclusions**

448 Latitudinal gradients in organic matter are the most important predictor of latitudinal changes
449 in species composition of oribatid mites across the spatial extent of Great Britain. These
450 changes partially drive variation in species richness and phylogenetic diversity but a
451 significant fraction of the variation observed in these metrics remained unexplained,
452 suggesting a potential role for unmeasured environmental heterogeneity at medium and
453 small scales. Despite small and medium scale heterogeneity, macroecological patterns in
454 this major group of soil animals are predictable by the climatic factors that control variation in
455 plant community structure and organic matter.

456

457 **Data accessibility**

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All Countryside Survey data used in this study are available from the NERC Environmental Information Data Centre. Data on oribatid mite records (<https://doi.org/10.5285/05cae9ae-86c4-4a2f-a628-6c7fd0882459>) are linked to soil properties (<https://doi.org/10.5285/9d1eada2-3f8b-4a7b-a9b0-a7a04d05ff72>) and vegetation (<https://doi.org/10.5285/07896bb2-7078-468c-b56d-fb8b41d47065>) by unique plots codes in each dataset. Access to spatial coordinates for plot locations in Countryside Survey is restricted to minimise risk of compromising future surveys and preserve anonymity of locations; plot locations at 10 × 10 km resolution, required to collate oribatid records at a hectad level and retrieve linked climatic data (Met Office), were therefore obtained under a bespoke data licence agreement between TC and the Centre for Ecology & Hydrology.

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661 ~~Data accessibility statement: The data supporting the results are available from the~~
662 ~~Environmental and Information Data Centre of NERC (UK). The spatial coordinates require a~~
663 ~~license agreement to be accessed.~~

665 Biosketch

666 **Tancredi Caruso** is a quantitative ecologist investigating the processes that structure
667 terrestrial biodiversity in space and time

668 **Ina Schaefer** is an evolutionary biologist with expertise on soil fauna

669 **Aidan M Keith** is an ecologist investigating land use and climate impacts on soil biology and
670 functional processes

671 **Frank Monson** is a soil animal taxonomist

672

673 All authors developed the concept of the paper; AMK and FM compiled and collected the
674 data, TC analysed the data, IS compiled and collected molecular data and constructed the
675 phylogenetic trees. All authors contributed substantially to the writing of the ms

676

Figure Legends

Figure 1 a) species richness and b) Chao's estimator of species richness for each 10 × 10 km plots. The maps were obtained via kriging interpolation at the hectad scale. Red represents high values, yellow low values. Species richness (a) displays a clear latitudinal gradient with richness increasing northward along the southwest-northeast direction whereas Chao's estimator (b) displays a patchy distribution suggesting the existence of hotspots if species richness.

Figure 2 Species-area relationships for the oribatid mites of Great Britain. The three fitted models (power law, semi-log, sigmoidal) all fit the data reasonably well but the AIC criterion clearly shows that the sigmoidal model provides the best fit (blue dotted line).

Figure 3 Principal Coordinates Analysis (PCoA) ordination of oribatid mites (a). The first axis (PCoA1) is a gradient that follow changes in mean Aggregate ~~vegetation~~Vegetation Class, with more organic vegetation types and woodland soil scoring on the positively site of PCoA1 and agricultural vegetation types (crops and grasslands) and arable soil scoring on the negatively side of PCoA1. This gradient is also correlated to organic matter, latitude, and precipitation, which are all positively correlated with PCoA1 ~~;- In fact,~~ a kriging interpolation of PCoA1 show a clear latitudinal gradient (b)

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Figure 4. Kriging interpolation of three metrics of Phylogenetic Diversity. The Faith's index (a) showed gradients that were very correlated to the same ones observed for plot species richness (Fig. 1a) while Mean Pairwise Distance (MPD) and Mean Nearest Taxon Distance (MNTD) mostly reflected longitudinal gradients; although MPD reaches the highest values in

the North and the East while MNTD seems more variable and ~~reaching-reaches~~ the highest value in the South-East.

Figure 5 Structural equation model linking latitude and abiotic parameters to oribatid mite species composition (PCoA1 of Fig. 3) and diversity (Faith Index PD and Mean Nearest Taxon Distance or MNTD of Fig. 4). Species richness was highly correlated to PD and was thus excluded, while MNTD and MPD returned similar results in this SEM and we selected MNTD, which provided the best fit. The model is supported by all metrics of global fit (Chi-square = 8.809 with 9 df and p-value of 0.185, CFI = 0.989 and RMSEA = 0.059). ~~Figures~~ Values beside ~~the~~ arrows are the path standardised coefficients. Black arrow stands for positive coefficient and gray arrows for negative coefficients. Paths statistically significant at p-value < 0.05 are in bold. All paths were statistically significant except for the direct effect of PD on MNTD and the direct effect of Latitude on PCoA 1. See also Supporting Information b for the full model output, including exact values of path coefficients, R-square values, standard deviations and statistical significance of parameter estimates. The model could account for 50, 16 and 5 % of variance in PCoA 1, PD and MNTD respectively. The model could also account for 55 and 17 % of variance in Organic Matter and Precipitation, respectively.