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Transcriptomic analysis of short-term 17A-ethynylestradiol exposure in two Californian sentinel fish species sardine (*Sardinops sagax*) and mackerel (*Scomber japonicus*)

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1 ***Transcriptomic analysis of short-term 17 α -ethynylestradiol exposure in two***
2 ***Californian sentinel fish species sardine (*Sardinops sagax*) and mackerel***
3 ***(*Scomber japonicus*)***

4
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25 17 α -ethynylestradiol; Genomic biomarkers
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ABSTRACT

Endocrine disrupting chemicals (EDCs) are substances which disrupt normal functioning of the endocrine system by interfering with hormone regulated physiological pathways. Aquatic environments provide the ultimate reservoir for many EDCs as they enter rivers and the ocean via effluent discharges and accumulate in sediments. One EDC widely dispersed in municipal wastewater effluent discharges is 17 α -ethynylestradiol (EE2), an orally bio-active estrogen used in almost all modern formulations of combined oral contraceptive pills and one of the most commonly prescribed medicines. As evidence of the health risks posed by EDCs mount, there is an urgent need to improve diagnostic tools for monitoring the effects of pollutants. As the cost of high throughput sequencing (HTS) diminishes, transcriptional profiling of an organism in response to EDC perturbation presents a cost-effective way of screening a wide range of endocrine responses. Coastal pelagic filter feeding fish species analyzed using HTS provide an excellent tool for EDC risk assessment in the marine environment. Unfortunately, there are limited genome sequence data and annotation for many of these species including Pacific sardine (*Sardinops sagax*) and chub mackerel (*Scomber japonicus*), which limits the utility of molecular tools such as HTS to study disruption of the endocrine system. In this study, we carried out RNA sequencing (RNAseq) of liver RNA harvested from wild sardine and mackerel exposed under laboratory conditions to EE2 and developed an analytical framework for transcriptomic analyses of species with limited genomic information. EE2 exposure altered expression patterns of key genes involved in important metabolic and physiological processes. The systems approach presented here provides a powerful tool for obtaining a comprehensive picture of endocrine disruption in aquatic organisms.

1. INTRODUCTION

The coast of Southern California is home to more than 20 million people. Servicing this high density urban strip are a series of wastewater treatment plants (WWTPs), which discharge more than a billion gallons of treated wastewater into the coastal marine environment on a daily basis [1]. Unfortunately, many contaminants are not filtered at the WWTPs and end up in the marine environment. These contaminants include a broad suite of compounds collectively referred to as endocrine disruptors (EDs) in concentrations that have been shown to disrupt the endocrine system. The derivative of estradiol 17 α -ethynylestradiol (EE2) has been a primary component of most modern formulations of combined oral contraceptive pills since the 1960s and qualifies as an ED due to its ability to bind and activate estrogen receptors (ERs) [2]. EE2 is a bio-active estrogen that is minimally metabolized and subsequently released into the environment via the urine and feces from individuals who take it as a medication. It has been detected in effluents from three WWTPs in Southern California at concentrations ranging from 31 to 43 ng/L (100 - 150 pM) [3] and is also detectable in surface waters in the USA (4 - 40 pM) [4] and in Europe (0.3 – 15 pM) [5, 6].

Active at very low concentration, EE2 has been shown to decrease reproductive potential by inhibiting gonad development in both male and female vertebrates, increasing ovarian follicle atresia and inducing complete sex reversal at concentrations ranging from 0.1 to 100 ng/L (0.3 - 340 pM) [7-11]. There is increasing global evidence that exposure to low levels of EDs can lead to disruption of hormone-mediated responses in fish [12-15].

The Pacific sardine (*Sardinops sagax*) and Pacific chub mackerel (*Scomber japonicus*) are coastal pelagic fishes that are major components of commercial fisheries as well as forage base for marine mammals and fishes in California [16, 17]. Sardines (*Sardinops spp.*) are found in most temperate upwelling systems of the world and have supported large commercial fisheries, though their populations are known to historically go through boom and bust periods of abundance, likely due to climatic oscillations and to a lesser degree by industrialized fishing in the past century [18-20]. Along the west coast of North America, Pacific sardines occur across gradients of temperature, salinity, and anthropogenic pollution from southern Canada into the Gulf of California. As filter feeders in a coastal environment, sardines can readily accumulate toxins from the environment and food they eat [21]. Mackerels (*Scomber spp.*) are found in warm and temperate coastal areas throughout the Atlantic, Pacific and Indian oceans. The Pacific chub mackerel is found throughout the Pacific and largely overlaps in distribution with the Pacific sardine where both species school together near the coast and interact with other schooling coastal pelagic species such as northern anchovy (*Engraulis mordax*) and jack mackerel (*Trachurus symmetricus*) [22]. These species are widely consumed by humans, owing to their rich omega-3 fatty acids composition [23] imparting both a distinct flavor profile as well as health benefits. Due to their low trophic diet, cosmopolitan distribution, affinity for coastal habitats, and importance as forage, they serve as ideal sentinel species to monitor the health of the coastal marine environment.

Management of coastal pelagic fish species requires a detailed understanding of the effects of pollutants on the physiology of these species. This need motivated our study to characterize the effect of environmentally relevant levels of EE2 on sardine and mackerel liver transcriptome, the primary organ impacted by pollution and main site of detoxification [24]. Despite their commercial and ecological importance, little is known about the sardine and mackerel genomes, a limiting factor for the use of conventional genomics tools (e.g. microarrays and Q-RT PCR) to study

96 endocrine disruption in these species. However, recent advances in DNA sequencing technologies
97 have enabled rapid production of transcriptomic data, which has enabled characterization of the
98 expression of specific genes in these species [25, 26]. Furthermore, these new technologies allow
99 for the development of biomarker fingerprints in sentinel species for different chemicals, such as
100 EE2.

101 In this manuscript, we describe **(1)** a novel RNA sequencing (RNAseq) pipeline using
102 scaffold transcriptomes from related species (**Supplemental Figure S1**) to streamline the data
103 analysis of HTS for sardine and mackerel exposed to EE2, and **(2)** examine the effects of EE2
104 exposure on these two sentinel species by performing a systems level analysis of their hepatic
105 transcriptome and provide common biomarkers to use as possible end points to predict the
106 estrogen-like contamination in environmental monitoring providing the most relevant Adverse
107 Outcome Pathways (AOP). This knowledge can help improve tests such as the USA EPA Office
108 of Prevention, Pesticides and Toxic Substances (OPPTS) Fish Acute Toxicity Test, Freshwater
109 and Marine and the Organization for Economic Co-operation and Development (OECD) Acute
110 Toxicity Test for Estuarine and Marine Organisms.

111 112 **2. MATERIALS & METHODS**

113 114 **2.1. Sampling of sardines and mackerels from Southern California coastal locations**

115 Sardine and mackerel were obtained from the live bait receiver operated by Everingham
116 Bros Inc. in Mission Bay, San Diego, California. Fish were allowed to acclimate to research tanks
117 at the NOAA Southwest Fisheries Science Center Experimental Aquarium facility under ambient
118 temperature (~17°C) flow through seawater conditions and fed a pelleted feed (Bio-Oregon) to
119 satiation. For exposure experiments (for both species), five males were exposed to 12.5 pM EE2
120 for five hours (environmentally relevant concentration in the range of surface water levels) and an
121 additional unexposed five males served as controls. Fish were euthanized with an overdose of
122 Tricaine Methanesulfonate (250mg/L) dissolved in seawater. Livers were harvested and flash-
123 frozen in liquid nitrogen at -196°C. These procedures followed an approved institutional IACUC
124 protocol and all animals were treated humanely. RNA was extracted from liver samples using
125 TRIzol reagent (Invitrogen, CA) and further purified using the RNeasy Mini kit with DNase to
126 remove DNA (Qiagen, Valencia, CA). RNA concentrations were determined at 260 nm using a
127 ND1000 (Nanodrop, Wilmington, DE). RNA was tested for structural integrity with the 6000 Nano
128 LabChip assay from Agilent, (Santa Clara, CA, USA). Only RNA samples with RIN scores > 7.0
129 were used for RNA-seq.

130 131 **2.2. High Throughput Sequencing (HTS)**

132 Libraries for RNA sequencing (RNAseq) were generated using Illumina TruSeq™ RNA
133 Sample Preparation Kit (Illumina, USA) in accordance with the manufacturer's recommendations
134 (**Figure 1A**). The library fragments were purified with AMPure XP system (Beckman Coulter,
135 USA) to select cDNA fragments of approximately 300 bp in length. DNA fragments with ligated
136 adaptor molecules on both ends were selectively enriched using Illumina PCR Primer Cocktail in
137 a 10-cycle PCR reaction. Products were purified using AMPure XP system and quantified using
138 the Agilent high sensitivity DNA assay on the Agilent Bioanalyzer 2100 system. Clustering was
139 performed on a cBot Cluster Generation System using TruSeq PE Cluster Kit v3-cBot-HS
140 (Illumina, USA). After cluster generation, the libraries were sequenced on an Illumina GAIIx

141 sequencer to a depth of approximately 5 million reads per sample and 101-bp single-end reads
142 were generated. Data was formatted to FASTQ format (with sequences and their associated Phred
143 quality scores) using CASAVA v1.8 (Illumina). RNA sequencing data has been submitted to the
144 Gene expression Omnibus with the accession number **GSE113780**.

145

146 **2.3. Transcript Level Analysis and Mapping to Reference Transcriptomes**

147 Low quality reads were filtered to exclude those most likely to represent sequencing errors.
148 Adaptor sequences were subsequently trimmed to generate clean full length reads. Only sequences
149 which had a Phred quality score of Q30 (which indicated 99.9% base call accuracy) were used for
150 downstream analysis.

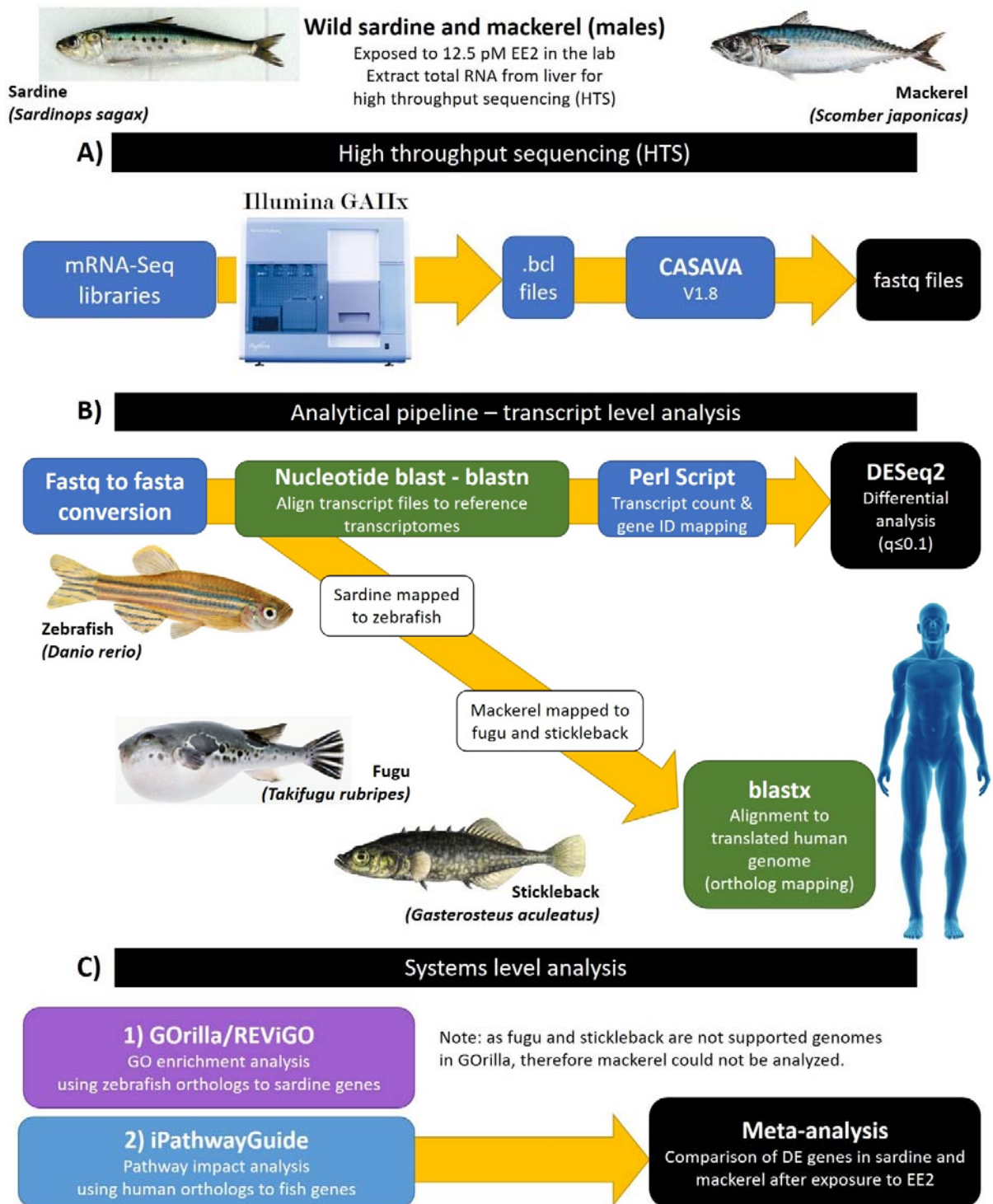
151 To analyze the mackerel and sardine RNAseq data, an analytical pipeline was developed
152 (**Figure 1B**) that starts with the conversion of fastq files to fasta format via a custom PERL script.
153 As comprehensive genomic sequence data for both sardine and mackerel are presently not
154 available, we exploited the phylogenetic similarity of sardine and zebrafish (**Supplemental Figure**
155 **S1**), and mapped sardine reads to the zebrafish (*Danio rerio*) Genome Reference Consortium
156 Zebrafish Build 10 (GRCz10) transcriptome using the Nucleotide-Nucleotide Basic Local
157 Alignment Search Tool (BLAST) nucleotide to nucleotide option (blastn) [27]. For mackerel, we
158 mapped reads to two reference transcriptomes, fugu (*Takifugu rubripes*) FUGU 4.0 and
159 stickleback (*Gasterosteus aculeatus*) *Gasterosteus aculeatus* BROAD S1. Fastq files were
160 converted to fasta files and blastn alignment was carried out against the reference transcriptomes
161 of interest formatted as fasta files.

162 After determining the threshold alignment length (see below), the reads were assembled
163 into transcript level expression summaries; the number of reads mapped to each gene or transcript,
164 with alignment length greater than the threshold alignment length, were summed, yielding a count
165 as a measure of transcript expression [28].

166 Each sample yielded a vector of counts or summed mapped reads, with length p equal to
167 the number of transcripts present. Combining the results from n samples, the data were formatted
168 as an $(n \times p)$ matrix of n transcripts and p samples. Log-log and square root plots were generated
169 from the data using custom R scripts [29]. The base of the logarithm was arbitrarily selected as 2
170 as the choice facilitated convenient preliminary data QC (**Supplemental Figure S2**).

171 In parallel, human ortholog refseq IDs were appended to the fish sequence reads. This was
172 achieved by aligning the reference fish transcriptome with the human proteome using blastx.
173 Utilizing the NCBI gene2refseq function, each human protein ID was linked with its corresponding
174 gene ID, such that the $(n \times p)$ matrix of transcript counts, with n number of reference transcriptome
175 genes contained the corresponding human gene IDs.

176 In order to infer differential gene expression with robust statistical power, we utilized
177 DESeq2, a method which tests for differential expression based on a model using negative binomial
178 distribution [30]. Transcript count data from DESeq2 analysis was ranked according to adjusted
179 p-value (or q-value), the smallest false discovery rate (FDR) at which a transcript is called
180 significant. FDR was calculated using the Benjamini-Hochberg multiple testing adjustment
181 procedure and the cut-off was set at $q \leq 0.1$.



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Figure 1: Workflow followed in this study starting with A) high throughput sequencing (HTS), B) transcript level analysis using a novel analytical pipeline, and C) systems level analysis of the GO enrichment and pathway impact.

186 **2.4. Systems Level Analysis**

187 Gene ontology (GO) enrichment analysis was performed on the sardine dataset mapped
 188 onto zebrafish genome as previously described [31, 32]. Briefly, GO terms were obtained using
 189 the online tool Gene Ontology enRIchment anaLysis and visuaLizAtion (GOrilla) (**Figure 1C-1**)
 190 [33]. GO terms were summarized using REduce & VISualize Gene Ontology (REViGO) that
 191 combines redundant terms into a single, representative term based on a simple clustering algorithm
 192 relying on semantic similarity measures [34]. As fugu and stickleback are not supported genomes
 193 in GOrilla at this point in time, mackerel mapped onto these references could not be analyzed using
 194 this tool.

195 Pathway impact analysis (**Figure 1C-2**) was performed using iPathwayGuide (Advaita
 196 Bioinformatics) [35], an approach that retrieves information from the Kyoto Encyclopedia of
 197 Genes and Genomes (KEGG) database for pathway analysis and diseases [36], from the Gene
 198 Ontology Consortium database for GO enrichment analysis [37], and from miRBase and
 199 TARGETSCAN databases for miRNA analysis [38, 39].

200 Area-proportional Venn diagrams were created using VENNY 2.1 [40].

201

202 **3. RESULTS**

203

204 **3.1. Cross species transcriptome mapping**

205 Due to the lack of reference genomes for sardine and mackerel, reads were mapped to the
 206 transcriptomes of phylogenetically similar fish species: sardine to the zebrafish transcriptome and
 207 mackerel to the fugu and stickleback transcriptomes (**Supplemental Table S1**). The mean of the
 208 total reads obtained with the sardine and mackerel samples were 4 and 3.7 million reads
 209 respectively. For sardine, about 1.4 million reads mapped to the zebrafish transcriptome, which
 210 represents 34% cross species mapping efficiency. For mackerel, about 1.5 and 3.5 million reads
 211 mapped to fugu and stickleback respectively, representing cross species mapping efficiencies of
 212 40.5% for fugu and 94% for stickleback.

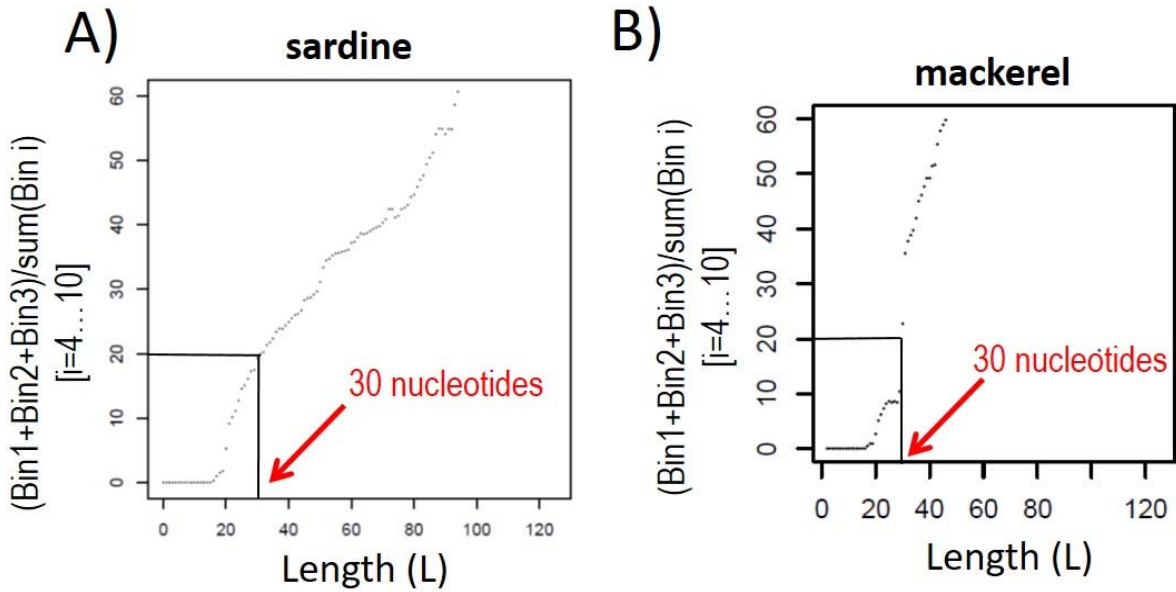
213

214 **3.2. Calculation of threshold alignment length to provide maximal alignment stringency**

215 Sequence read lengths were 101bp. In order to determine the minimum alignment length
 216 that provided the highest number of unique reads, we followed a read binning strategy so that a
 217 given read was placed in bin 0 if there was no alignment with length $\geq L$. When there were n
 218 alignments $\geq L$, the read was placed into bin $\min(n, 10)$. Utilizing this binning approach, based on
 219 alignment lengths a comma-separated values (csv) file was created with row numbers equal to L
 220 + 1 (L = threshold alignment length) and columns corresponding to bins 0 to 10. A series of QC
 221 plots were generated with the alignment length plotted on the x-axis, and the individual and
 222 fraction of combination of bins on the y-axis. From the plot of $(\text{bin1} + \text{bin2} + \text{bin3}) / (\text{bin4} + \dots + \text{bin10})$
 223 on the y-axis and an alignment length L on the x-axis, which was essentially the fraction of unique
 224 (U) to non-unique ($1-U$) versus alignment length (L), it was deduced for sardine and mackerel with
 225 the $U/(1-U)$ ratio of 20, that the optimal threshold alignment length was 30 nucleotides (nt), i.e.
 226 95% of the aligned reads were unique with an alignment length of 30 nt (**Figures 2A and 2B**
 227 **respectively**).

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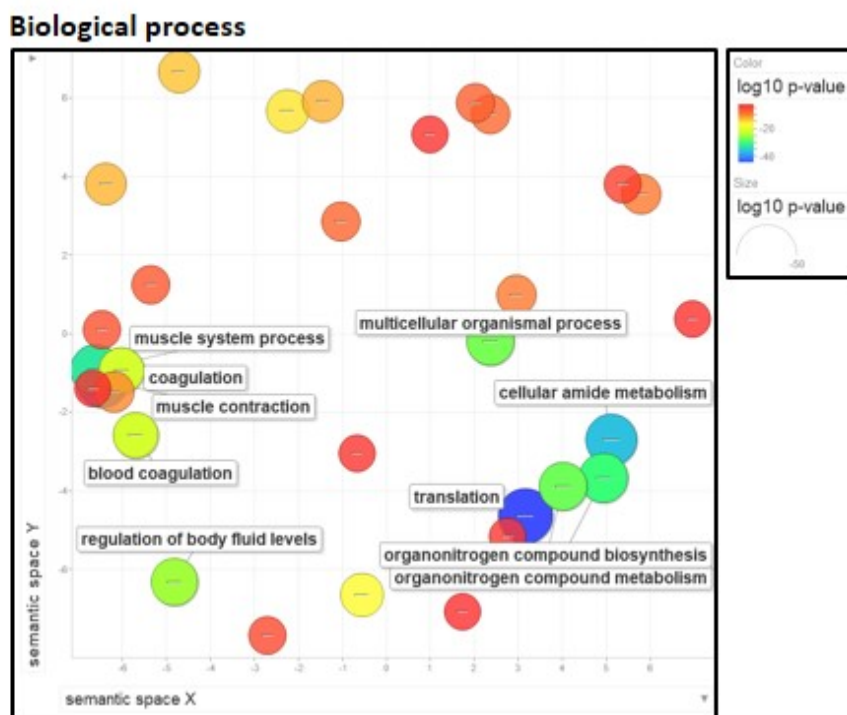


230
 231 **Figure 2:** Threshold alignment length ($n=30$) for sardine (A) and mackerel (B). Alignment lengths
 232 plotted on the x-axis (labeled “L”), and the individual and fraction ratio of bin combinations on
 233 the y-axis. For sardine and mackerel with the $U/(1-U)$ ratio of 20, the optimal threshold alignment
 234 length is 30 nucleotides, i.e. 95% of aligned reads are unique (see red arrows).

235

236 **3.3. GO analysis of fish genes**

237 Gene level analysis performed with DESeq2 revealed 4,779 and 2,211 DE hepatic mRNAs
 238 in sardine and mackerel exposed to EE2 respectively. For systems level analysis, we first entered
 239 the sardine/zebrafish DE gene list in GOrilla since annotations for zebrafish are available. GOrilla
 240 GO enrichment analysis showed that translation, muscle contraction, cellular amide metabolism,
 241 muscle system, organonitrogen compound metabolism and biosynthesis, multicellular organismal
 242 process, regulation of body fluid levels and blood coagulation were some of the most enriched BP
 243 terms after exposure to EE2 in the liver of sardines (**Figure 3, deep blue to light green bubbles,**
 244 **and Supplemental Table S2).**
 245



246 **Figure 3: EE2 exposed sardine: GOrilla GO enrichment analysis - Biological process.** The
 247 REViGO scatterplots show cluster members in a 2-dimensional space obtained by applying multi-
 248 dimensional scaling to a matrix of the GO terms' semantic similarities (the axes have no intrinsic
 249 meaning). Bubble color indicates the $-\log_{10} p\text{-value}$ (legend in upper right-hand corner, blue and
 250 green bubbles are GO terms with greater significant p-values than the orange and red bubbles),
 251 the size indicates the GO term frequency in the GO database (bubbles of more general terms are
 252 bigger), and the proximity on the plot reflects the semantic similarity.
 253

254

255 **3.4. Pathway impact analysis (iPathwayGuide)**

256 Since the zebrafish annotations are not as detailed as human annotations [32], we entered
257 the human orthologs of the sardine/zebrafish and mackerel/fugu DE genes in iPathwayGuide for a
258 content richer GO and pathway impact analysis. Impact analysis uses two types of evidence: i) the
259 overrepresentation of differentially expressed (DE) genes in a given pathway and ii) the
260 perturbation of that pathway computed by propagating the measured expression changes across
261 the pathway topology. Volcano plots revealed a greater dynamic range in the sardine dataset as
262 compared to mackerel (**Supplemental Figure S3**).

263 DE analysis in sardine exposed to EE2 identified 3,918 DE genes out of a total of 5,321
264 genes with measured expression (**Supplemental Figure S3A**). Significant differences between the
265 EE2 exposed and control sardines in metabolic, steroid biosynthesis and cancer pathways emerged
266 (**Supplemental Table S3 – Biological Pathways**) as well as enrichment in several signaling
267 pathways such as Fc epsilon RI, cAMP, oxytocin and VEGF. Additionally, complement &
268 coagulation cascades and focal adhesion biological pathways were also significantly perturbed.

269 The GO analysis revealed that multicellular organismal (**Supplemental Table S3 –**
270 **Biological Process**), muscle, steroid, metabolic (includes alcohol and cholesterol metabolic
271 processes) related biological processes were highly enriched after exposure to EE2. Response to
272 chemical and drug, negative regulation of peptidase and endopeptidase activity as well as positive
273 regulation of immune system process were also highly represented BP terms in this analysis. Other
274 general BP GO terms were also listed including regulation of biological quality, ion transport,
275 developmental process and system process. Together these results suggest that exposure to EE2 in
276 sardine had a strong impact on transcripts related to metabolism, steroid biosynthesis, several
277 enzymes and signaling pathways as well as the immune system, response to drug/chemical, and
278 cancer pathways.

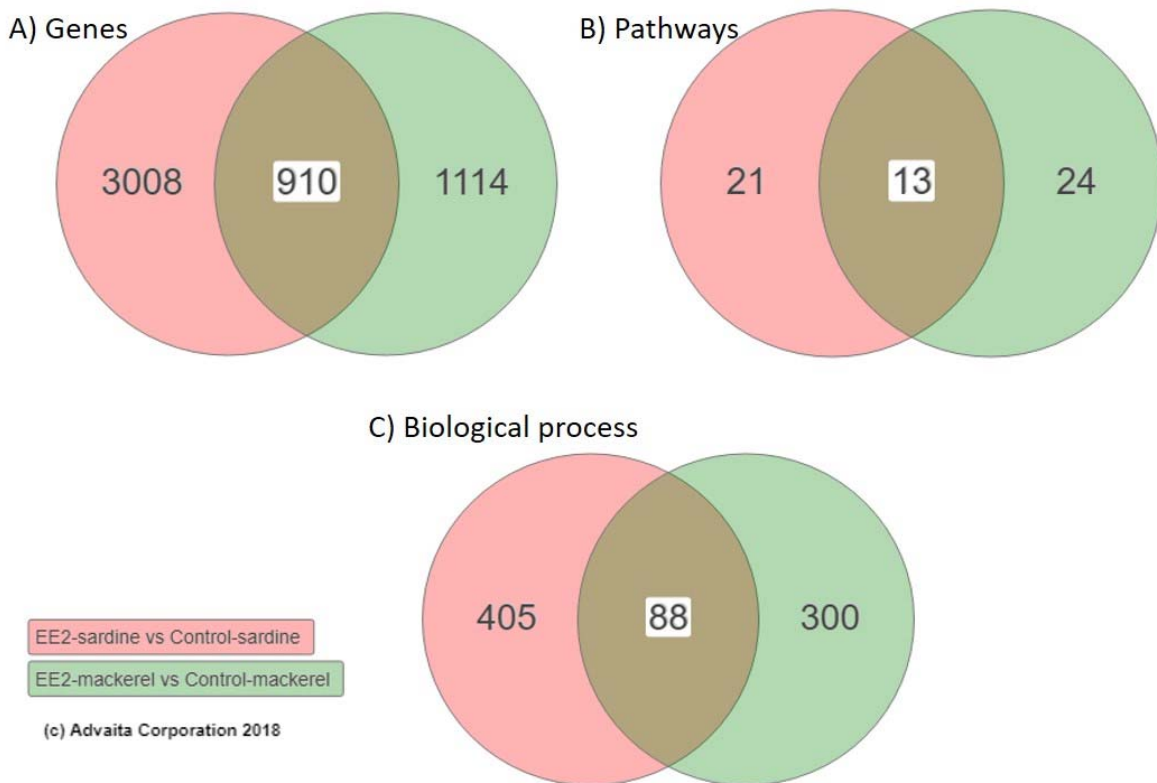
279 Exposure to EE2 in mackerel perturbed 2,024 DE genes out of 5,300 genes with measured
280 expression (**Supplemental Figure S3B**). The most enriched biological pathways in EE2 exposed
281 mackerel are fat digestion & absorption, metabolic, steroid biosynthesis and cancer pathways
282 (**Supplemental Table S4 – Biological Pathways**) as well as enrichment in several signaling
283 pathways such as PI3K-Akt and Rap1. ECM-receptor interaction, necroptosis, EGFR tyrosine
284 kinase inhibitor resistance and proteasome pathways were also highly significant in this analysis.

285 Several BP terms related to metabolic (**Supplemental Table S4 – Biological Process**) and
286 steroid biosynthetic processes were highly enriched in the GO analysis. More specifically,
287 enriched metabolic processes can be sub-divided into alcohol, cholesterol and lipid related
288 processes. Together, these results suggest that exposure to EE2 in mackerel disrupts digestion,
289 steroid biosynthesis, several metabolic processes and signaling pathways, and enhances cancer
290 pathways.

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294 **3.5. Meta-analysis (iPathwayGuide)**

295 In order to compare the effects of EE2 exposure on the hepatic transcriptomes of sardine
 296 and mackerel we carried out a meta-analysis. This revealed that 910 DE genes (**Figure 4A**), 13
 297 enriched pathways (**Figure 4B and Supplemental Table S5**) and 88 BP (**Figure 4C**) are shared
 298 by sardine and mackerel. Additionally, this analysis highlighted that each species has a unique
 299 signature after EE2 exposure; 3,008 DE genes, 21 pathways and 405 BP are unique to sardines
 300 (**pink in panels A, B, C**) while 1,114 DE genes, 24 pathways and 300 BP are unique to mackerel
 301 (**green in panels A, B, C**).
 302



303 **Figure 4:** Venn diagrams from iPathwayGuide. **A)** Intersection between liver DE genes expressed
 304 in sardine and mackerel in response to EE2 exposure; 910 DE genes are commonly expressed in
 305 both species while 3,008 and 1,114 DE genes are unique to sardine and mackerel respectively. **B)**
 306 Intersection between enriched pathways in sardine and mackerel; 13 pathways are common to
 307 both species while 21 and 24 pathways are unique to sardine and mackerel respectively. **C)**
 308 Intersection between enriched Biological Process (BP) in sardine and mackerel; 88 BP are
 309 commonly expressed in both species while 405 and 300 BP are unique to sardine and mackerel
 310 respectively.
 311

312

313 4. DISCUSSION

314

315 *4.1. Development of a novel analytical pipeline for sardine and mackerel*

316 Despite the commercial and ecological importance of sardine and mackerel, very little
 317 genomic information is available at the present time. The first goal of this study was to develop a
 318 novel analytical pipeline to facilitate **1)** mapping reads to reference transcriptomes using blastn
 319 when no reference transcriptome is available for a given species, **2)** calculation of threshold
 320 alignment length to provide maximal alignment stringency, **3)** assembly of reads with transcript-
 321 level counts or expression summaries, **4)** sequence quality control, **5)** annotation of sequence reads
 322 with human GO identifiers, **6)** data normalization (to adjust the read counts for variance in
 323 sequencing depths between different samples), differential expression analysis and FDR
 324 estimation using DEseq2, **7)** generation of significant DE transcript lists, and **8)** GO and pathway
 325 impact analysis.

326 Based on fish phylogenetic relationships and available reference transcriptomes, we
 327 aligned sardine fasta files to the zebrafish transcriptome using blastn, which provided a cross
 328 species mapping efficiency of 34%. In the case of mackerel, the fasta files were aligned to both
 329 the fugu and stickleback transcriptomes (note that a draft genome sequence has been released for
 330 stickleback which is a closer evolutionary relative of mackerel than fugu). The number of mapped
 331 sequence tags was considerably higher when stickleback was chosen as the reference (94% cross
 332 species mapping efficiency) compared with fugu (40.5%). However, fugu was used as the
 333 reference (for all downstream analyses) as this genome is currently better annotated than
 334 stickleback [41, 42]. Therefore, zebrafish and fugu transcriptomes served as scaffolds for mapping
 335 purposes in our study.

336 Analysis of the BLAST data output revealed multiple hits for each single query read of
 337 varying alignment length to the cognate transcripts. As sequences were being compared across fish
 338 species, full length alignment was not expected and consequently many of the small alignment
 339 length hits were not unique or meaningful. It was therefore necessary to parse out smaller
 340 sequences and determine the optimal alignment length. This was calculated for both fish species
 341 so that the only transcript retained was the one whose alignment length was the maximum and
 342 greater than the threshold alignment length for each read.

343

344 *4.2. Systems level analysis of transcriptomic changes using zebrafish annotations*

345 Utilizing the online tools GOrilla and REViGO, we performed GO enrichment analysis
 346 using sardine mapped to zebrafish genes for the sardine dataset. This analysis showed that
 347 translation (including redundant GO terms peptide and amide biosynthetic process), muscle system
 348 & contraction, and cellular amide metabolism were some of the most enriched BP terms after
 349 exposure to EE2 in the liver of sardines, followed by organonitrogen compound metabolism and
 350 biosynthesis, regulation of body fluid levels and blood coagulation (fibrin clot
 351 formation/fibrinolysis, negative regulation of wound healing, hemostasis and blood coagulation).
 352 The enriched BP terms include genes that surprisingly mapped to GO BP terms related to muscle
 353 system and contraction. This may reflect physiological differences between fish and mammals.
 354 The histology of the fish liver for example differs from its mammalian counterpart in that there is
 355 a far less tendency of the hepatocytes to form distinct cords or lobules, and the typical portal triads
 356 are not obvious [43]. Many of these muscle system & contraction mRNAs are associated with
 357 disease signatures (e.g. mouse phenotype database MP: 0003141; cardiac fibrosis and MP:

358 0005608; cardiac interstitial fibrosis). EE2 exposure has been linked with hepatotoxicity and
359 fibrosis [44] so it is conceivable that these mRNA signatures of fibrosis in other tissues are being
360 detected in this analysis. The liver plays a central role in the clotting process, and acute and chronic
361 liver diseases have been linked with coagulation disorders [45]. Our GO analysis highlighted that
362 EE2 exposure impacted “blood coagulation” as mentioned above, reinforcing that the liver plays
363 a role in the production of clotting factors as well as red blood cell production [46, 47].

365 **4.3. Systems level analysis of transcriptomic changes using human orthologs**

366 Another challenge is the fact that zebrafish and fugu have limited gene annotations when
367 compared to human, further limiting the resolution of the analysis [32]. To remedy this short-
368 coming, human refseq IDs and GO terms were appended to the fish sequence reads by aligning
369 the reference fish transcriptomes with the human proteome using blastx. As 69% of zebrafish genes
370 have at least one human ortholog [48], this approach allows a more sensitive comparative
371 genomics based analysis of transcriptomic changes in these sentinel species. Utilizing the NCBI
372 gene2refseq function, each human protein ID was linked with its corresponding gene ID. However,
373 fish and mammalian species exhibit considerable diversity, each possessing unique biological
374 characteristics. By “humanizing” the fish dataset, species specific genes are lost, *i.e.* vitellogenin,
375 and artefactual data may result if the orthologs are not carefully mapped. However, in the absence
376 of annotated genomes and GO identifiers for the majority fish species, this approach represents a
377 valid analytical alternative that provides functional characterization of the fish transcriptome.
378 Additionally, the BLAST program has an internal p-value threshold, ensuring that the homology
379 between the human and reference fish proteome reported is deemed significant. Furthermore,
380 information obtained about impacted human pathways also provide insights on the potential effects
381 to human health of a particular contaminant.

383 **4.3.1. Effects of EE2 on sardine**

384 The most obvious result from our study is that exposure to 12.5 pM EE2 greatly affected
385 sardine transcripts related to metabolism, including carbon metabolism (KEGG 01200, **Figures**
386 **S4-S5**), glycolysis and gluconeogenesis (KEGG 00010, **Figures S6-S7**). In particular, the main
387 metabolic perturbation we observed was on the biosynthesis of steroid hormone (KEGG 00140,
388 **Figures S8-S9**). This is consistent with previous studies conducted in turbot [49] and zebrafish
389 [50], and this is the first report to our knowledge of this hormonal disruption induced by EE2 in
390 sardines. Given that steroid hormones are derived from cholesterol and that our GO analysis
391 highlighted that cholesterol biosynthetic and metabolic processes were significantly impacted by
392 exposure to EE2, it is consistent with our previous observation. In rat, EE2 has been shown to
393 decrease testosterone production via a reduction of the cAMP pathway during steroidogenesis [51].
394 This is interesting as the cAMP signaling pathway (KEGG 04024, **Figures S10-S11**) was one of
395 the most enriched Biological Process in our GO analysis. The cAMP pathway regulates pivotal
396 physiologic processes including metabolism, secretion, calcium homeostasis, muscle contraction,
397 cell fate and gene transcription [35]. cAMP is one of the most common and universal second
398 messengers and its formation is promoted by adenylyl cyclase (AC) activation by ligands including
399 hormones. Estrogenic hormones have been shown to have a non-genomic action that activates
400 second-messenger signaling systems such as cAMP, explaining how steroid hormones can alter
401 the expression of genes [52]. To our knowledge, no studies conducted in fish have examined the
402 impact of EE2 on the cAMP signaling pathway.

403 Several other signaling pathways were enriched upon EE2 exposure, including VEGF,
404 FcεRI, and oxytocin signaling pathways. The VEGF signaling pathway (KEGG 04370, **Figures**
405 **S12-S13**) is a crucial signal transducer in both physiologic and pathologic angiogenesis [35]. Its
406 activation leads to proliferation and migration of endothelial cells. Recent studies in human and
407 primate have shown that estrogen regulates VEGF expression [53, 54], however the effects of EE2
408 on VEGF signaling pathway in fish have not been well characterized to date.

409 Activation of the FcεRI signaling pathway (KEGG 04664, **Figures S14-S15**) in mast cells
410 induces the release of preformed granules containing histamines, proteoglycans (PGs),
411 prostaglandins and cytokines that contribute to inflammatory responses [35]. Interestingly, many
412 PGs have been shown to be key macromolecules that contribute to the development of various
413 types of cancer including proliferation, adhesion, angiogenesis and metastasis [55]. Our analysis
414 revealed that the pathway “proteoglycans in cancer” was highly enriched in EE2-treated sardines
415 (KEGG 05205, **Figures S16-S17**), suggesting that exposure to EE2 can disrupt many cancer-
416 associated pathways, such as FcεRI and proteoglycan pathways. Minimal research has been carried
417 out on the effects of EE2 on the FcεRI signaling pathway and PGs, highlighting the need for further
418 examination of this subject.

419 Additionally, the FcεRI signaling pathway plays a role in innate immunity as mast cells
420 have been shown to play a critical role in defense against bacterial, viral and fungal pathogens [56].
421 Mast cells are activated during innate immune responses by multiple mechanisms, including well-
422 established responses to complement components. We showed here that the “complement &
423 coagulation cascades” pathway (KEGG 04610, **Figures S18-S19**) was highly perturbed by EE2
424 (consistent with our previous GO analysis with GOrilla using zebrafish annotations), suggesting
425 that EE2 disrupted the innate immunity of exposed sardines. The effect of combined oral
426 contraceptive pills containing EE2 on coagulation parameters has been shown in human [57] but
427 more investigation is needed to understand the effect of EE2 on sardine’s innate immunity.
428

429 **4.3.2. Effects of EE2 on mackerel**

430 In mackerel, the most enriched biological pathway upon EE2 exposure was “fat digestion
431 & absorption” (KEGG 04975, **Figures S20-S21**). Given that fat is an important energy source
432 from food, this emphasizes that EE2 even at low concentration can severely impair major basic
433 pathways. The second most disturbed pathway in EE2-treated mackerel was “chemical
434 carcinogenesis” (KEGG 05204, **Figures S22-S23**). In fact, exposure to environmental chemical
435 carcinogens may account for the majority of human cancers [35] by interacting with a variety of
436 genotoxic and non-genotoxic mechanisms. Genotoxic mechanisms can generate DNA adducts
437 (segments of DNA bound to a cancer-causing chemical) either by direct impact on the DNA or
438 indirectly by affecting metabolism. Non-genotoxic carcinogens induce inflammation,
439 immunosuppression, formation of reactive oxygen species, activation of receptors and epigenetic
440 silencing [35]. Our data suggest that EE2 acts as a carcinogen in mackerel potentially by altering
441 signal-transduction pathways resulting in hypermutability, genomic instability, loss of
442 proliferation control and resistance to apoptosis. This is consistent with previous studies conducted
443 in rats [58], in hamsters [59] and zebrafish [60].

444 Moreover, EE2 also perturbed “central carbon metabolism in cancer”, another pathway
445 that is central to malignant transformation of cells (KEGG 05230, **Figures S24-S25**); as Otto
446 Warburg discovered, a major shift in cellular metabolism is required in cancer cells in order to
447 support growth and survival. This shift results in the consumption of large amounts of glucose,

448 maintaining high rates of glycolysis and converting glucose into lactic acid even under normal
449 oxygen concentrations [61, 62].

450 Several signaling pathways were highly enriched upon EE2 treatment in mackerel,
451 including PI3K-Akt and Rap1 signaling pathways. The PI3K-Akt signaling pathway (KEGG
452 04151, **Figures S26-S27**) is activated by many types of cellular stimuli or toxic insults and
453 regulates fundamental cellular functions such as transcription, translation, proliferation, growth,
454 and survival [35]. PI3K catalyzes the production of phosphatidylinositol-3,4,5-triphosphate (PIP3)
455 at the cell membrane, a second messenger that helps to activate Akt. Once active, Akt can control
456 key cellular processes by phosphorylating substrates involved in apoptosis, protein synthesis,
457 metabolism, and cell cycle. Our data suggest that exposure to EE2 affects fundamental processes
458 by regulating second messenger in mackerel, consistent with our previous conclusion that EE2
459 regulates cAMP and translation in sardines. Interestingly, estrogens have been shown to induce
460 VEGF-A expression in rat pituitary tumor cells through PI3K-Akt dependent signaling pathway,
461 independently of activation of estrogen receptors [63]. However, the effects of EE2 on the
462 mackerel's PI3K-Akt pathway remain undefined.

463 The Rap1 signaling pathway (KEGG 04015, **Figures S28-S29**) controls diverse processes
464 such as cell adhesion, cell-cell junction formation and cell polarity [35]. Rap1 plays a dominant
465 role in the control of cell-cell and cell-matrix interactions by regulating the function of integrins
466 and other adhesion molecules in various cell types. Our analysis also revealed that the “ECM-
467 receptor interaction” biological pathway was perturbed by EE2 exposure in mackerels (KEGG
468 04512, **Figures S30-S31**). The ECM serves an important role in the maintenance of cell and tissue
469 structure and function. Specific interactions between cells and the ECM are mediated by
470 transmembrane molecules, mainly integrins and also proteoglycans, CD36, or other cell-surface-
471 associated components [35]. These interactions lead to a direct or indirect control of cellular
472 activities such as adhesion, migration, differentiation, proliferation, and apoptosis. Together, this
473 suggests that EE2 disrupted cell-cell and cell-matrix mechanisms by perturbing Rap1 signaling
474 and ECM-receptor interaction, in agreement with previous *in vitro* studies [64]. *In vivo* data also
475 suggest that estrogens can exert direct regulatory effects on endothelial cells by increasing surface
476 expression of integrins and enhancing integrin-mediated signaling [65]. Since integrins have an
477 important role in mediating endothelial cell attachment, migration and differentiation, the increase
478 in integrin expression and function induced by estrogens may promote neovascularization and
479 vessel repair. Moreover, our data show that EE2 diminished the expression of three collagen genes
480 (*COL1A1*, *COL1A2* and *COL6A3*). This is consistent with Hansen *et al.* [66] who concluded that
481 administration of EE2 in women suppressed synthesis of collagen in tendon in response to exercise.
482 Our data also show an increase in expression of *ITGB1* in response to EE2. Castillo-Briceno *et al.*
483 [67] have previously demonstrated that EE2 induces liver and kidney toxicity and injuries in fish
484 via upregulation of ECM-related molecules, including *ITGB1*.

485 Our analysis also revealed that the “necroptosis” biological pathway was perturbed in EE2
486 exposed mackerel (KEGG 04217, **Figures S32-S33**). Necroptosis is a programmed form of
487 necrosis. It can be initiated by different stimuli, such as tumor necrosis factor (TNF), TNF-related
488 apoptosis-inducing ligand (TRAIL), Fas ligand (FasL), interferon (IFN), LPS, viral DNA or RNA,
489 DNA-damage agent and requires the kinase activity of receptor-interacting protein 1 (RIPK1) and
490 RIPK3 [35]. Its execution involves ROS generation, calcium overload, the opening of the
491 mitochondrial permeability transition pore, mitochondrial fission, inflammatory response and
492 chromatinolysis. Necroptosis participates in pathogenesis of many diseases, including

493 neurological diseases, retinal disorders, acute kidney injury, inflammatory diseases and microbial
494 infections. This cell death process is evolutionarily conserved and a *caspase-3* gene has been
495 identified in zebrafish [68] and sea bass [69], retaining the motifs that are functionally important,
496 such as the active site and the cleavage site at the aspartic residue [70]. The examination of the
497 effects of EE2 exposure on necroptosis in fish is in its infancy and further investigation is needed
498 to fully understand its impact in mackerel.

499 Exposure to EE2 perturbed the “biosynthesis of amino acids” pathway as well (KEGG
500 01230, **Figures S34-S35**). This KEGG module is the most conserved one in the KEGG MODULE
501 database and is found in almost all completely sequenced genomes [35]. Although human
502 nutritional requirements are well-characterized, little is known of the specific nutrient
503 (protein/amino acid, lipid, carbohydrate, mineral and vitamin) requirements of cultivated fish
504 species; even in zebrafish, a commonly used model, this information is almost entirely lacking
505 [71]. Requirements for amino acids should be delineated in feeding studies with different strains
506 and life stages before the effect of a given contaminant can fully be assessed on this pathway.
507 Nevertheless, our data are consistent with a previous study reporting that the most sensitive
508 biological process to EE2 exposure was amino acid glycosylation [72]; in fact, the authors argued
509 that the genes relating to amino acid glycosylation may be a more sensitive set of biomarkers than
510 the current single standard biomarker of vitellogenin.

511 512 **4.4. Meta-analysis (Common pathways and biomarkers in sardine and mackerel and** 513 **extrapolation to humans).**

514 We exploited iPathwayGuide to perform a meta-analysis of the effects of EE2 on the
515 hepatic transcriptome of the sardine and mackerel. Meta-analysis revealed that several DE genes,
516 biological pathways and BP are commonly affected in both species whereas others are unique to
517 sardine or mackerel. “Carbon metabolism (**Figure S36**), steroid hormone biosynthesis (**Figure**
518 **S37**), glycolysis & gluconeogenesis (**Figure S38**), Rap1 signaling pathway (**Figure S39**) and
519 chemical carcinogenesis” (**Figure S40**) are amongst the 13 perturbed biological pathways shared
520 in both species and discussed in the section above. “Glyoxylate & dicarboxylate metabolism, drug
521 metabolism (other enzymes), retinol metabolism, metabolic pathways, arginine & proline
522 metabolism, steroid biosynthesis, pathways in cancer and pentose phosphate pathway” are also
523 commonly disrupted pathways in both species.

524 Taken together, the meta-analysis reveals that exposure to 12.5 pM EE2 has a significant
525 impact on basic biological responses in both fish species, impacting essential metabolic processes
526 and biological pathways. These results highlight that EE2 deregulates signaling pathways that are
527 involved in the development and progression of cancer as well as the endocrine and immune
528 systems. The literature on estrogen-induced carcinogenesis (and contaminants with estrogenic
529 properties) is substantial [73-75] yet the exact mechanisms and pathways involved are still under
530 investigation. One established mechanism involves direct genomic action by activation of
531 estrogen-receptor signal transduction pathways associated with increased cell proliferation and
532 inhibition of apoptosis [76-78] via perturbation of genes controlling cell cycle [76] and other
533 tumor-promoting factors such as VEGF [53, 54, 79]. Our findings are consistent with these studies;
534 in sardine, the VEGF signaling pathway was significantly perturbed after EE2 exposure and 4
535 Biological Process GO terms related to cell cycle were enriched, including “regulation of G1/S
536 transition of mitotic cell cycle and positive regulation of cell cycle G2/M phase transition”
537 (**Supplemental Figure S41**).

538 Estrogens and estrogenic contaminants can also interfere with cell cycle regulation by
539 upregulating mitogen-activated protein kinases and secondary messengers, such as cAMP, via
540 non-genomic action [77, 78, 80]. Accordingly, we determined that the “cAMP and PI3K-Akt
541 signaling pathways” were significantly impacted in EE2-treated sardines and mackerels
542 respectively. Estrogen metabolites have been shown to have genotoxic (formation of DNA adducts
543 and oxidative DNA damage), mutagenic, transforming, and carcinogenic effects [81, 82]. This is
544 consistent with the enrichment of the “chemical carcinogenesis” pathway in mackerels exposed to
545 EE2 in our study.

546 Much effort has been expended in search of alternate model for estrogen-related human
547 cancer research. Lam *et al.* have demonstrated that molecular conservation of estrogen-
548 responsiveness between zebrafish and human cancer cell lines makes the zebrafish a great model
549 and established common mechanisms for estrogen-induced carcinogenesis [83]; from an
550 evolutionary perspective, their findings suggest that estrogen regulation of cell cycle is perhaps
551 one of the earliest forms of steroidal-receptor controlled cellular processes. Upon further
552 characterization of the sardine and mackerel transcriptomes, these two species have the potential
553 for future use as additional cancer research models.

554 In summary the rationale for exploring liver tissue in this study stemmed from the fact that
555 the liver is the primary organ for metabolism and detoxification. The motivation for carrying out
556 EE2 exposure experiments was to facilitate extrapolation of toxicological insights from fish
557 species to human and provide insights into ocean and human health. Our work has uncovered
558 common biomarkers of low level EE2 exposure which provides a basis for improved risk
559 assessment practices based on ‘Big Omics Data’. Strategies have been developed by others to
560 derive AOPs (for example, domoic acid exposure) [84]. Perkins and co-workers employed a
561 reverse engineering strategy based on network inference to uncover key nodes (i.e., genes, proteins,
562 metabolites) impacted by chemical perturbation [85, 86]. Kramer and colleagues successfully
563 demonstrated that models can predict potential population impact, once the data are transformed
564 into the prediction of an adverse outcome of demographic significance (e.g., reproduction) at the
565 organism level [87]. Celander *et al.* developed a comparative genomics approach with CYP19
566 aromatase, and predicted potential aromatase toxicity of new compounds across a multiplicity of
567 species [88]. Systematic organization of Big Data Science into AOP frameworks likely can
568 improve regulatory decision-making through greater integration and more meaningful exploitation
569 of mechanistic data. However, in order to develop a useful knowledgebase that encompasses
570 toxicological contexts of concern to human health risk assessment, novel approaches such as the
571 RNAseq pipeline we described must be developed in accordance with AOP core principles.

572 573 **5. CONCLUSION**

574 The primary goal of this study was to develop a novel analytical pipeline to elucidate the
575 effects of environmental levels of EE2 on the hepatic transcriptome in sardine and mackerel, two
576 species that do not have a well-characterized genome. Using scaffold transcriptomes from related
577 species (zebrafish for sardine, and fugu for mackerel), we were able to streamline the HTS data
578 analysis and examine the effects of low dose EE2 exposure on these species by performing a
579 systems level analysis of their hepatic transcriptome. We determined that even at low doses, EE2
580 disrupted basic biological processes and pathways, leading to a signature of metabolic, hormonal
581 and immune dysfunction, as well as carcinogenesis in exposed fish. This work uncovered common
582 dysregulated genes and biological pathways in both species providing biomarkers for EE2

583 exposure in the marine environment, and facilitating monitoring of ocean environmental health.
584 As many of these processes and pathways are well conserved from an evolutionary perspective,
585 and as our analytical pipeline projected the fish genes onto their human orthologs we can infer the
586 impact that EE2 exposure could have on human-health. Our results call for further examination of
587 the effects of estrogenic contaminants on fish and human health at concentrations that represent
588 relevant levels of environmental exposure.
589
590

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592

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